Authors thanks the reviewers and the Editor for their time and constructive comments and suggestions, which we believe have improved the manuscript by making it more clearly and consistent.

Comments from Referee 1 (Report #1)

The manuscript still needs a substantial grammatical revision. In the first revision, this reviewer recommended the authors to hire an English-proofreading expert. My sensation is that they did not follow the advice. Major issues with English remain (e.g. wrong verb tenses, incorrect spelling).

An English-proofreading expert has finally revised the entire manuscript.

Several sections still need to be improved (see detailed comments in the text of the uploaded PDF file).

The authors are often too vague in relation to some approaches or methodologies, which are not clearly explained. More accuracy is also needed when presenting some of the results.

Authors addressed the main concerns from the reviews. Detailed comments can be seen at the end of this document: Authors replied to all the comments/revisions/suggestions posted by the reviewer #1.

Comments from Referee 2 (Report #2)

The study investigates the evolution and main factors determining LULC changes in Portugal during the period 1990-2012, providing a description of how it has affected the RUI’s extension and evolution. An assessment of the burnt areas during the study period is also undertaken. The study presents useful information in order to interpret the landscape changes that have taken place in recent decades in Portugal, helping to put in context the observed fire activity. In my opinion the authors have satisfactorily addressed most of the concerns raised by the previous referees, leading to a substantially improved manuscript with a clearer structure and conclusions well sustained by the results presented. The topic of the study is well suited for the aims of the NHESS journal, and has a good fit within the topics covered by the targeted special issue With this regard, I recommend its acceptance. I give in the following a few suggestions to the authors for their consideration.

Minor comments/suggestions

The provision of RUI maps is an important outcome of the study, as highlighted by the authors. With this regard, I think that many interested readers would greatly benefit from these data to be open for public usage so others can directly benefit from your results and build new knowledge based on them. This would give much more visibility to the paper and would surely boost the number of references to the paper. Of course, I understand that concerns may exist with openly sharing all the data, so this is just a suggestion for the authors to consider the possibility of opening some of the data presented, and in particular the CORINE-based RUI maps, in view of their inherent value.

Authors thanks the reviewer for this suggestion and they provided the four maps of RUI (i.e. RUI 1990, 2000, 2006 and 2012) as shapefile. These have been uploaded as Supplement, zip file named "RUI_Maps" containing the shapfiles.

I have some minor concerns that I will next enumerate for your consideration. I am not myself a native English speaker (thus often making mistakes), but I find some minor typos, too long sentences or some awkward expressions (e.g. P2L7 “is far to be homogeneously distributed”) and a few grammatical inconsistencies that still occur within the text after the author’s revision. This is nothing too serious, but
that needs to be revised. I have just highlighted a few issues in the abstract in the paragraphs below. Therefore I would recommend a second text review before publication.

An English-proofreading expert has finally revised the entire manuscript.

The analysis of burned area evolution is very interesting. As a mere suggestion, I think that it would also add value to the author’s findings to put the observed trends of burned area within a wider context (national/regional), taking into account the discussion the results presented in other studies (see e.g. the trends for Portugal and other Mediterranean countries described by Turco et al. 2016), highlighting the importance of the new data presented in this paper to interpret the observed fire trends. In my opinion the last paragraph of the Discussion (P16 L31-34) needs some reference supporting this claim regarding future climate projections. Here are a few suggestions relevant for the specific case of Portugal: Parente et al 2018; Sousa et. al 2015; Bedia et al 2014,

Authors added the suggestion references and revised the paragraph of Discussion

**Suggestions regarding paper Figures**

The number of figures is elevated. I don’t see any problem with it, as long as all figures display relevant information, that aids to understand the methodology and to interpret the results. As a suggestion, it would perhaps be possible to merge Figures 1 and 2 into a single double-panel figure, without loss of readability. It would also help to better compare side-by-side the population density and main land cover types.

Authors followed the suggestion of the reviewer of merging Fig. 1 and Fig. 2

Regarding Fig 8, I would have opted by a 2D barchart (either using side-by-side or stacked categories) rather than a 3D graph, that can lead in my humble opinion to somewhat difficult visualization. But of course I fully respect the author’s choice.

Authors believe that a 3D chart provide a better visualization of the results
Global assessment of rural-urban interface in Portugal related to land cover changes

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Abstract. The rural-urban interface (RUI), known as the area were structures and other human development meet or intermingle with wildland and rural area, is at present a central focus of wild-fire policy and its mapping is crucial for wildfire management and impact. In the Mediterranean basin, the majority of fires are caused by humans; humans cause the vast majority of fires and fire risk is particularly high in the proximity of infrastructures and of rural/wildland areas. RUI’s extension changes under the pressure of environmental and anthropogenic factors, such as urban growth, fragmentation of rural areas, deforestation and, more in general, land use/land cover changes (LULCC). As other Mediterranean countries, Portugal experienced significant LULCC in the last decades in response to migration, rural abandonment, ageing of population; and trends associated to the high socioeconomic development. In the present study, we analysed the LULCC occurred in this country in the 1990—2012 period with the main objective of investigating how these changes affected RUI’s evolution. Moreover, we performed a qualitative and quantitative characterization of burnt areas within the RUI in relation the LULCC observed changes was performed. Obtained results disclose important LULCC and reveal which their spatial distribution, which is far from uniform within the territory. A significant increase in artificial surfaces was registered nearby the main metropolitan communities of the northwest and littoral-central and southern regions, whilst the abandonment of agricultural land nearby the inland urban areas led to an increase of uncultivated semi-natural and forest areas. Within agricultural areas, heterogeneous patches suffered the greatest changes and were the main contributors to the increase of urban areas; moreover, this land cover class, together with forests, resulted to be highly affected by wildfires in terms of burnt area. Finally, from this analysis and during the investigated period, it appears that during the investigated period RUI increased in Portugal more than two thirds, while the total burnt area decreased one third; nevertheless, burnt area within RUI doubled, which emphasizes the significance of RUI monitoring for land and fire managers.

1 Introduction

Mediterranean area region is particularly affected by wildfires, mainly as consequence of its type of climate and vegetation cover fire proneness (Pellizzaro et al., 2012; Amraoui et al., 2015). In the European Mediterranean countries, fire incidence has dramatically increased in the last decades and the average total annual burnt area (hereafter, BA) has quadrupled since the 60’s (San-Miguel-Ayanz et al., 2012), mainly due to changes in climate and land use (Moreira et al., 2011; Ferreira-Leite et al., 2016). Portugal stands out from this group of countries since it counts the highest number of wildfires and has been the third most affected country in terms of BA in the last three decades (Pereira et al., 2014; San-Miguel-Ayanz et al., 2016). On average, about 95% of wildfires with known causes in Europe during the period 1995 to 2010 (corresponding to about 70% of the total number of recorded
events) were associated to human activities (Ganteaume et al., 2013). While only a small percentage of wildfires (e.g. 1% in Portugal and 5% in Spain) were naturally caused by lightnings (Mateus and Fernandes, 2014; Vilar et al., 2016). Wildfires have long been considered a dynamic ecological factor and an efficient agricultural and landscape management tool. In the last decades, the increasing increase of the human population along with the consequent expanding of the urban area and land use/land cover changes (LULCC), made the interfaces between the wildland and the human assets vulnerable to wildfires more abundant and critical. Therefore, wildfires but more recently they are increasingly considered a hazard (Bond and Keeley, 2005; Fernandes and Botelho, 2003; Hardy, 2005; Moreno and Oechel, 2012; Pyke et al., 2010; Van Wagendonk, 2007), which has motivated governments to implement measures for fires prevention, monitoring and mapping.

In this regard, a key factor in this regards is to study the investigation of the spatial and temporal distribution patterns of wildfires into a given area-region is absolutely fundamental in this regard. Normally, this pattern is not randomly distributed, which, being influenced by the surrounding socio-economic and environmental factors, is far to be homogeneously distributed. For example, clusters of wildfires were highlighted in Portugal, a clustered pattern was discovered for Portugal, with wildfires hot spots concentrated in the north-west and center regions, while the southern area-region presented lower densities of wildfires (Pereira et al., 2015; Tonini et al., 2017). Urban sprawl also affects the spatial-temporal pattern of these hazardous events, and makes it difficult to define a boundary between human developments and rural areas, in order to better protect this interface where wildfires are more likely to occur. The modern urban landscape in Europe has a typical star-shaped spatial pattern (Antrop, 2000) with wedge of unchanged countryside persisting between lobes of urban development (Antrop, 2004). In this context, populations and activities described either as “rural” or “urban” are closely linked and their distinction is often arbitrary (Tacoli, 1998). In the Iberian Peninsula, the diffuse urban expansion along radiating access roads connecting the center to commercial/industrial zones and isolated housing, as well as the growth of peri-urban centers in previously rural areas, leaves gaps in the suburban/exurban space of urban agglomerations (Trigal, 2010). Here the expansion and reconfiguration of urban and metropolitan areas comprise the processes of sub-urbanisation and peri-urbanisation, mainly of coastal regions and, in the interior, of agricultural areas (Trigal, 2010). Land use changes are at the origin of landscape patterns and dynamics and have a strong influence on forest fires (Silva et al., 2011). On the one hand, each vegetated land cover type, such as agricultural, natural and semi-natural vegetation cover, has a specific fire proneness depending on the differences in vegetation structure, moisture content and fuel load composition (Barros and Pereira, 2014; Oliveira et al., 2014; Pereira et al., 2014). Further, fire occurrence affects landscape dynamics by changing the vegetation structure and the soil processes according to the fire adaptation of each ecosystem (Viedma, 2008; Pausas et al., 2009).

Population and urban areas significantly increased in Europe during the late XX century, which helps to understand the rapid land cover and land use/land cover changes (LULCC) (Noronha Vaz et al., 2012). In the European Mediterranean countries, LULCC are mainly caused by the increasing migration to urban centers, at the cost of the abandonment of rural areas, and by the expansion of costal tourism (Alodos et al., 2004; Tedim et al., 2016). One consequence of this process is the urbanization, defined as the process involving the transformation of rural and natural landscape into urban and industrial areas, caused by the interaction of very different factors and largely influenced by communication, accessibility and mobility needs (Antrop, 2000). The dynamic conversion among rural and urban spaces can become extremely complex (Strubelt and Deutschland, 2001): urbanization generates the centralization of certain area by changing the land use, population density, economical activities and transportation network. This complex process is at the origin of the abandonment of remote rural areas with poor accessibility, which allows the
expansion of wild low vegetation and forest (Antrop, 2004). Specifically, the abandonment of low-intensity agricultural lands and grazing practices caused the increase of forest cover and scrubland vegetation (Poyatos et al., 2003; Millington et al., 2007).

Significant LULCC occurred in Portugal in the recent period. The urbanization of coastal areas in the country occurred in concomitance with the abandonment of agricultural land in marginal areas and seemed to prevail between 1990 and 2000 while, in the later period (2000–2006), predominated the intensification of agriculture in areas where irrigation was available was a predominant process (Diogo and Koomen, 2012; Van Doorn and Bakker, 2007). Pereira et al. (2014) observed that among the southern European countries, in the period 2000–2006 Portugal registered the highest rates of land use change in the period 2000–2006 period, marked by a significant increase in artificial surfaces and Sclerophyllous vegetation and a decrease in forest area and natural grasslands, because of rural abandonment, urbanization and wildfires.

The interface between the wildland and the urban space, called Wildland-Urban Interface (WUI), has been deeply investigated by researchers and fire managers in the last decades. The United States Department of Agriculture (USDA), defined the WUI as the area “where humans and their development meet or intermix with wildland fuels” (Stewart et al., 2007); here in this area, fires can spread readily among vegetation fuels and urban structures. Anthropogenic features, such as the distance to roads and houses, negatively have a negative influence on the probability of forest fire occurrence, while the population density positively affects it (Haight et al., 2004; Radeloff et al., 2005; Stewart et al., 2007; Hammer et al., 2009; Lampin-Maillet et al., 2010; Conedera et al., 2015). There are strong evidences that the expansion of the urban and WUI area increased the fire density and related risk (Fox et al., 2015; Gallardo et al., 2016; Lampin-Maillet et al., 2010; Viedma et al., 2015). There are strong evidences that the expansion of the urban and WUI area increased the fire density and related risk (Fox et al., 2015; Gallardo et al., 2016; Lampin-Maillet et al., 2010; Viedma et al., 2015).

Researchers developed several geospatial models for defining and mapping the WUI, taking into account all the above-mentioned factors. In Europe, Lampin-Maillet et al., (2010) proposed an approach tested in southern France and based on the combination of four types of buildings configuration and three classes of vegetation structure—tested in southern France. Following this model, Bouillon et al., (2012) developed WUImap, a software tool for mapping WUI in the Mediterranean region successfully applied in southeastern France, eastern Spain and Sardinia, in Italy. In the Alpine context, geospatial approaches to map the WUI were developed in Switzerland by Conedera et al., (2015) and in France by Fox et al., (2015). Pellizzaro et al., (2012b) characterized and mapped the WUI in Sardinia, Italy, using temporal steps of about 10 years from 1954 and 2008, and found an increase of the WUI’s extension. Most of these researches were performed locally, at house-spatial-scale or for small regions within each country, suggesting the need for a homogeneous methodology applicable at national scale. In this regard, Amato et al., (2018) prosed a new procedure based on Multilayer Perceptron and Fuzzy Set Theory to map of the Rural-Urban-Interface for the entire Portugal; this approach allowed to develop continuous non-categorical maps expressing the possibility of being part of this interface in a future scenario. Other studies focused on additional other aspects related to the WUI, namely hazard/risk, vulnerability and fire risk management in Spain—(Badia et al., 2011; Galiana-Martin et al., 2011; Herrero-Corral et al., 2012). As well as fuel and fire modelling in France (Cohen et al., 2003; Pugnet et al., 2013). The majority of these researches were developed performed...
locally, and performed at house spatial scale or for small regions within each country, suggesting the need for a homogeneous methodology applicable to national or continental scale.

The active rural-urban conversion processes and the associated landscape changes, largely documented in the European context, encouraged to reconsider the WUI concept. In this respect, recent studies defined the Rural-Urban Interface (RUI) as an alternative to the WUI, to highlight the importance of including the rural areas, and identified the RUI as the most fire prone areas in Mediterranean countries (Badia-Perpinyà and Pallares-Barbera, 2006; Catry et al., 2010; Moreira et al., 2009). In the present study, the authors investigated the RUI in Portugal—____with two main objectives: (i) was to analyze changes in land use/land cover occurred in this country in the period 1990–2012 period; and—(ii) to assess their impact on RUI’s evolution. Moreover, the authors performed a qualitative and quantitative characterization of the burnt areas within the RUI in relation to the LULCC was performed. Finally, this research provides a first attempt to a map of the RUI’s extension and evolution at national level in the last twenty years (from 1990 to 2012) for the entire continental Portugal.

2 Study area

Continental Portugal (c.a. 90,000 km²) is located in the southwest of the Iberian Peninsula, bathed by the Atlantic Ocean on the south and west coast and bounded by Spain at north and east. According to the census data, the Portuguese population was about 10.6 million in 2011 and decreased to 10.3 million in 2017; its distribution displays a much higher density in the northwestern and southern coastal areas as well as around the major cities (Fig. 1). The Tagus river divides the country into two regions of approximately the same area but very different in terms of several biophysical and human drivers (Fig. 1). The north region is characterized by a temperate climate, with dry and warm summer, altitude ranging from sea level to about 2000 m and mean watercourse density of about 0.65 km/km². According to CORINE Land Cover (hereafter, CLC) inventory 2012 (EEA, 2016), 54% of north’s area is covered by forests and scrublands, 40% is used for agriculture and about 5% is occupied by artificial surfaces (Fig. 2). The region south of Tagus river is characterized by temperate climate with dry and hot summer, low altitude range (between sea level and about 1000 m), and mean watercourse density of about 0.58 km/km². According to CLC 2012 inventory, agricultural areas occupy 56% of this territory is occupied by agricultural areas, 40% by forest and semi-natural areas 40%, and only 2% by...
Figure 1 — (Left) Population density at parish level (INE, 2012) in the mainland Portugal, with the location of the major cities. NUTS refers to regions according to Nomenclature of Territorial Units for Statistics level II (Eurostat, 2017; Santos, 2014). (Right)
3 Data and methodology

Wildfires digital data for the period 1990 – 2015 came from the National Mapped Burned Area dataset (NMBA), provided by the Portuguese Institute of Conservation of Nature and Forests (ICNF). for the period 1990 – 2015, The Institute of Agronomy (Instituto Superior de Agronomia, ISA) produced the official database in the earlier years (1975 to 1989) and finally the official National fire database covers the 1975–2015 period and comprises about 49,000 fire events for a total BA of 4,430,000 ha (Oliveira et al., 2012; Barros and Pereira, 2014). The annual fire perimeter maps resulted from a semi-automatic supervised image classification procedure (Gorte, 1999) performed with the classification and regression trees algorithm of Breiman et al., (1984) on late summer-autumn Thematic Mapper/Enhanced Landsat satellite imagery. Technological improvements of satellite sensors allowed acquiring and processing over-time higher-resolution images with an increasing accuracy-resolution from the initial 30 hectares to 5 hectares after 1984, and even higher after 2005. To ensure the accuracy of the dataset, results were compared against field statistics gathered on the ground by the National Forest Authority and by the National Civil Protection Authority. For each fire record, the dataset comprises the BA (perimeter map) and the year of occurrence.

Land use/land cover’s information came from the CLC inventory provided by European Environment Agency (EEA). CLC is delivered as cartographic product, both in raster (i.e. a regular grid of cells) and in-vector-shapefile (as point, line and polygons) format. The minimum cartographic unit is of 25 ha (500 by 500 m) with a minimum geometric accuracy of 100 m and a thematic accuracy over 85% (EEA, 1994). CLC nomenclature is a three-level hierarchical classification system with 44 classes at the third and most detailed level (Table 1). The five more general classes for the first level are the following: Artificial Surfaces (AS), Agricultural Areas (AA), Forest and Semi-Natural Areas (FSNA), Wetlands, and Water bodies. CLC inventories are currently available for four periods (1990, 2000, 2006, and 2012) with a minimum time consistency of plus/minus one year. CLC was already been used for land-use change and urban dynamics studies in Portugal (Noronha Vaz et al., 2012; Pereira et al., 2014).

In order to identify and detail the major habitats/plant communities/vegetation types corresponding to each CLC class in Portugal, we employed the Soil Use and Occupancy Chart (Carta de Uso e Ocupação do Solo, COS) provided by the Portuguese Directorate-General of the Territory (Direção-Geral do Território, DGT). DGT is the national public body responsible for pursuing public policies for land use and town planning. COS is a national cartographic product with a minimum cartographic unit of 1 ha. We compared CLC2006 with COS2007v2.0 because these are the closest inventories within the study investigated period. In addition, COS2007v2.0 presents improvement from the thematic and geometric point of view (Sarmento et al., 2016): it includes 225 classes (32 more than the initial version) at the most detailed level, distributed over 5 hierarchical levels.

In the present study, the four CLC inventories were employed to analyze LULCC at different levels and to map RUI at different periods, according to the methodology described below.

The first analysis consisted in investigating LULCC within the entire study period (i.e. from 1990 to 2012). This allowed elaborating the map of changes showing the transitions among the five more general classes (CLC first level hierarchy) and, more in detail, to quantify gained and lost areas with reference to both the first and the second level hierarchy of CLC. Moreover, the difference between gains and losses within each class divided by the total area covered by the specific class in the later period, i.e. representing the net change, was computed and the result expressed in percentage.
RUI was then mapped for each period (1990, 2000, 2006, 2012) using a geospatial approach designed to extract the area of intersection between a buffer around the AS and the area resulting from the sum of the FSNA plus the Heterogeneous Agricultural Areas (HAA, a sub-level of AA). Different buffer widths from 100 m to 2 km were tested. Finally, we adopted a buffer width of 1 km, corresponding to two times the spatial resolution of CLC inventories. This value is in line with the values applied in other countries for WUI mapping (Radeloff et al., 2005; Vilar et al., 2016) and, at the same time, is large enough to avoid bias in the results, due to the CLC spatial resolution. The other agricultural areas (i.e. arable lands, permanent crops, pastures) were not included in the RUI definition since these vegetated land covers are usually well managed, mostly irrigated and frequently constitute an obstruction to fire spread. Similarly, San-Miguel-Ayanz et al., (2012) suggested that only HAA have to be considered in the definition and quantification of the RUI in Portugal, together with FSNA.

The geocomputation, which allowed mapping the production of the RUI's maps, was performed under ArcGIS™ software environment. Namely, the geoprocessing workflows was implemented into a Model Builder (Fig. 23), a specific application used to create, edit and manage models, meant as workflows that string together sequences of geoprocessing tools (e.g. selection, buffer, intersect), feeding the output of one tool into another tool as input (i.e. the raster or vector digital data). Finally, we analyzed how each land cover class (in respect of considering the third level hierarchy of CLC) was affected by wildfires in terms of BA for each investigated period within the RUI. To this end, polygons defining the BA registered at each CLC-year plus/minus one year (1989–1991, 1999–2001, 2005–2007, 2011–2013) were merged together and the resulting BA polygons were clipped over the corresponding RUI map. The resulting outputs, representing the BA within the RUI cumulated over three years around each investigated period, were finally overlapped with the CLC source map.

<table>
<thead>
<tr>
<th>Level 1</th>
<th>Level 2</th>
<th>Level 3</th>
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<tbody>
<tr>
<td>1 Artificial surfaces</td>
<td>11 Urban fabric</td>
<td>111 Continuous urban fabric</td>
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<tr>
<td></td>
<td></td>
<td>112 Discontinuous urban fabric</td>
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<tr>
<td></td>
<td>12 Industrial, commercial and transport units</td>
<td>121 Industrial or commercial units</td>
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<td></td>
<td></td>
<td>122 Road and rail networks and associated land</td>
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<td></td>
<td></td>
<td>123 Port areas</td>
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<tr>
<td></td>
<td></td>
<td>124 Airports</td>
</tr>
<tr>
<td>13 Mine, dump and construction sites</td>
<td>131 Mineral extraction sites</td>
<td></td>
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<tr>
<td></td>
<td>132 Dump sites</td>
<td></td>
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<tr>
<td></td>
<td>133 Construction sites</td>
<td></td>
</tr>
<tr>
<td>14 Artificial, non-agricultural vegetated areas</td>
<td>141 Green urban areas</td>
<td></td>
</tr>
<tr>
<td></td>
<td>142 Sport and leisure facilities</td>
<td></td>
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<tr>
<td>2 Agricultural areas</td>
<td>21 Arable land</td>
<td>211 Non-irrigated arable land</td>
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<td></td>
<td></td>
<td>212 Permanently irrigated land</td>
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<td></td>
<td></td>
<td>213 Rice fields</td>
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<td></td>
<td>22 Permanent crops</td>
<td>221 Vineyards</td>
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<td></td>
<td></td>
<td>222 Fruit trees and berry plantations</td>
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<td>223 Olive groves</td>
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<td>23 Pastures</td>
<td>231 Pastures</td>
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<tr>
<td>24 Heterogeneous agricultural areas</td>
<td>241 Annual crops associated with permanent crops</td>
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<tr>
<td></td>
<td>242 Complex cultivation patterns</td>
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<td></td>
<td>243 Land principally occupied by agriculture, with significant areas of natural vegetation</td>
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</tr>
<tr>
<td></td>
<td>244 Agro-forestry areas</td>
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<td>3 Forest and semi natural areas</td>
<td>31 Forests</td>
<td>311 Broad-leaved forest</td>
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<td>312 Coniferous forest</td>
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<tr>
<td></td>
<td></td>
<td>313 Mixed forest</td>
</tr>
<tr>
<td></td>
<td>32 Scrub and/or herbaceous vegetation associations</td>
<td>321 Natural grasslands</td>
</tr>
<tr>
<td></td>
<td></td>
<td>322 Moors and heathland</td>
</tr>
<tr>
<td></td>
<td></td>
<td>323 Sclerophyllous vegetation</td>
</tr>
<tr>
<td></td>
<td></td>
<td>324 Transitional woodland-shrub</td>
</tr>
</tbody>
</table>
8

33 Open spaces with little or no vegetation
331 Beaches, dunes, sands
332 Bare rocks
333 Sparsely vegetated areas
334 Burnt areas
335 Glaciers and perpetual snow

4 Wetlands
41 Inland wetlands
411 Inland marshes
412 Peat bogs
42 Maritime wetlands
421 Salt marshes
422 Salines
423 Intertidal flats

5 Water bodies
51 Inland waters
511 Water courses
512 Water bodies
52 Marine waters
521 Coastal lagoons
522 Estuaries
523 Sea and ocean

<table>
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<tr>
<th>CLC</th>
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<th>AS</th>
<th>Buffer 1 km</th>
<th>AS+</th>
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<tbody>
<tr>
<td>FSNA + HAA</td>
<td>Intersect</td>
<td>Rural Urban Interface</td>
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</table>

Table 1: CORINE Land Cover (CLC) nomenclature. Numbers on the left represent the CLC code. (Source: EEA, 2016)

4 Results

4.1 LULC change analysis

The analysis of main changes between CLC1990 and CLC2012 in the area occupied by the first level hierarchy classes between CLC1990 and CLC2012 allowed to visualize the main transitions occurred within the entire investigated period (Fig. 34) and thus to have an overview of the LULCC occurred in Portugal. It resulted that major main transitions occurred between vegetated areas (i.e. AA and/or FSNA) to artificial surfaces AS (AS) and between FSNA and AA in both directions. AS increased mainly nearby the main metropolitan communities of the northwest and littoral coastal central and southern regions. A transition from vegetated areas (AA and FSNA) to AS is also visible in center north and is probably due to the intensification of the main road network to connect the emergent inhabited rural area. The conversion from FSNA to AA and vice-versa appeared to be an active and dynamic process prevailing in the southern half of the country, but it was revealed also in the inner northern region. -Figure 45 shows the areas gained and lost for each CLC first-level class and the net percentage of changes, computed relatively to the total area of each class in the later land cover class. The main changes in terms of surface area were registered by AS, which increased 165×10^3 ha, and AA, which decreased 184×10^3 ha, but in terms of net percentage of change the increase of AS was about 50%, while AA decreased only 4.4%. The two classes which manly contributed to the increase in AS were AA, with 110×10^3 ha, and FSNA, with 50×10^3 ha.
Figure 34 – Map of land cover/land use transition from 1990 and 2012, evaluated considering the first level hierarchy of CLC 1990 and CLC 2012.

Figure 35 – Area lost and gained from 1990 to 2012 for each CORINE land cover class, considering the first level hierarchy. Net percentage changes were computed relatively to the total area of each class in the later land cover.
A more detailed analysis was carried out to investigate the changes from 1990 to 2012 occurred within classes considering of the second level of hierarchy from 1990 to 2012. Figure 56 shows that the majority of the CLC classes (level 2, Table 1) displayed important net changes in terms of relative gains and losses compared with values for the same classes in the later period. Scrub and/or herbaceous vegetation associations (code 32) registered a net gain of about 520 ×10³ ha (+24%), while the Forest area (code 31) decreased about 460 ×10³ ha (-23%). Arable land (code 21) was the only AA registering an important negative net change of -225 ×10³ ha (-20%). Among AS, Urban fabric (code 11) significantly increased 110 ×10³ ha (45%), and, in terms of net percentage of change by class, all the other three AS sub-levels, including Industrial/commercial and transport unit (code 12), Mine/dump and construction sites (code 13), Artificial/non-agricultural vegetated areas (code 14), increased more than half.

![Figure 56](image-url)  
**Figure 56** – Area lost and gained from 1990 to 2012 for each CORINE land cover classes, considering the second level hierarchy. Net percentage changes were computed relatively to the total area of each class in the later land cover.

The bar graph of the contributions to net changes in the AS sub-levels (Fig. 62) shows that Urban fabric (orange bars), which includes buildings, roads and artificially surfaced areas, grew mainly at the expense almost exclusive of HAA (code 24). On the other hand, the increase of Industrial commercial and transport (blues bars) was mainly due to the decrease of Forests (code 31), HAA (code 24) and Scrub and/or herbaceous vegetation associations (code 32).
5 Spatial distribution and characterization of burnt areas

Almost all the CLC third-level classes belonging to FSNA (code 3) were affected by wildfires, in terms of burned BA area (Table 2), with the Transitional woodland-shrub (code 324) and Mixed forest (code 313) as the first and second more damaged classes. This trend was similar in all the four investigated frame-periods, as highlighted in Figure 7, where the same results are expressed in percentage for each CLC classes considering only the areas within the RUI, as the ratio of BA over the total BA for the entire frame period, for each CLC classes considering only the areas within the RUI. The peak of BA in Transitional woodland-shrub (code 324) equals to about 43 × 10^3 ha in the 2005–2007 period, compared with about 15 × 10^3 ha in 2011–2013, 14 × 10^3 ha in 1999–2001 and 6 × 10^3 ha in 1989–1991. It also emerges that the three over four sub-levels of HAA (code 243, 241, 242) are highly affected by wildfires, thus confirming the need of including HAA in the RUI’s definition.

<table>
<thead>
<tr>
<th>CLC code</th>
<th>CLC classes</th>
<th>BA within RUI (ha)</th>
</tr>
</thead>
<tbody>
<tr>
<td>324</td>
<td>Transitional woodland-shrub</td>
<td>6086.31</td>
</tr>
<tr>
<td>313</td>
<td>Mixed forest</td>
<td>4368.85</td>
</tr>
<tr>
<td>312</td>
<td>Coniferous forest</td>
<td>3104.72</td>
</tr>
<tr>
<td>322</td>
<td>Moors and heathland</td>
<td>1835.03</td>
</tr>
<tr>
<td>243*</td>
<td>Land principally occupied by agriculture, with significant areas of natural vegetation</td>
<td>1535.36</td>
</tr>
<tr>
<td>------</td>
<td>--------------------------------------------------------</td>
<td>-----------</td>
</tr>
<tr>
<td>311</td>
<td>Broad-leaved forest</td>
<td>1144.72</td>
</tr>
<tr>
<td>241*</td>
<td>Annual crops associated with permanent crops</td>
<td>698.75</td>
</tr>
<tr>
<td>242*</td>
<td>Complex cultivation patterns</td>
<td>677.27</td>
</tr>
<tr>
<td>321</td>
<td>Natural grasslands</td>
<td>638.39</td>
</tr>
<tr>
<td>323</td>
<td>Sclerophyllous vegetation</td>
<td>562.68</td>
</tr>
<tr>
<td>334</td>
<td>Burnt areas</td>
<td>396.87</td>
</tr>
<tr>
<td>333</td>
<td>Sparsely vegetated areas</td>
<td>321.85</td>
</tr>
<tr>
<td>332</td>
<td>Bare rocks</td>
<td>123.84</td>
</tr>
<tr>
<td>244*</td>
<td>Agro-forestry areas</td>
<td>17.25</td>
</tr>
<tr>
<td>331</td>
<td>Beaches, dunes, sands</td>
<td>5.22</td>
</tr>
</tbody>
</table>

*a* Heterogeneous agricultural areas

Table 2 – Classes of land use, as defined by the third level hierarchy of the CORINE Land Cover (CLC), affected by wildfires expressed in terms of Burned Area (BA) within the Rural Urban Area-Interface (RUI) during three investigated frame periods.

Figure 78 – Classes of land use, as defined by the third level hierarchy of the CORINE Land Cover, affected by wildfires expressed in percentage as the ratio of Burnt Area (BA) affecting each CLC classes over the total BA for each frame period.

4.3 RUI map
In the present study, the classes FSNA and HAA were considered to describe the flammable rural area, which, intermingling with the urban area, defines the RUI. Thus, RUI maps arose from the zone of intersection between the sum of HAA plus FSNA and an enhanced area around AS (Fig. 3). The result was a zone of interface evolving in space and in time due to LULCC (Fig. 89). Analyzing the period 1990–2012, the increase of RUI was more active in the north-west and along the coast, where the transition from HAA to Urban fabric was particularly intense. This evolution was mainly due to the urban growth and to the intensification of the road network.

Figure 89 – Maps of the rural-urban interface (RUI) in Portugal estimated for the different periods of investigation, and corresponding to the available CORINE Land Cover inventories (1990, 2000, 2006, and 2012)

The total size of the RUI, the fraction of BA within the RUI (BAR) and the total BA (TBA) were computed (Fig. 94). It resulted that RUI increased from about 780×10^3 ha in 1990 up to about 1310×10^3 ha in 2012 following a power-law (RUI=776310.year^0.1686, R^2=0.99). Moreover, we computed the contribution both to the RUI and to the BA within the RUI (BAR) of each CLC class that make up the RUI, namely HAA (code 24), Forests (code 31), Scrub and/or herbaceous vegetation associations (code 32), and Open spaces with little or no vegetation (code 33). Results can be summarized as follows.
(Fig.- 9-10): (a) the relative contribution of the those four CLC classes to the RUI increases in time at approximately the same rate in each period almost equally; (b) HAA (code code 24) is the CLC class with the largest area within the RUI (~50%); (c) in terms of relative BA within the RUI (BAR), the most affected class is the Scrub and/or herbaceous vegetation associations (code 32), followed by Forest (code code 31), -HAA (code code 24) and then Open spaces with little or no vegetation (code 33). The total extent of the BA values fluctuated, with a maximum in 1990 (about 500×10^3 ha) followed by 2006 (~460×10^3 ha), while in 2000 and 2012 its value was lower and equal to about 310×10^3 ha. The portion of BA included within the RUI (RBA), expressed as percentage over the total BA, tends to increase in time, passing from 4% and 7% in 1990 and 2000 to 15% and 14% in 2006 and 2012, respectively.

5 Discussion

The LULCC analysis performed in the present study indicates that from 1990 to 2012 AS (code code 1) globally increased in Portugal about 50% (Fig. 45). This growing process is in good agreement with previous findings of other authors (Marques et al., 2014; Meneses et al., 2014; Oliveira et al., 2017; Tavares et al., 2012). Moreover, the present study confirms that the urban growth process in Portugal (quantified as changes in AS) was principally firstly caused by the transition from HAA (code code 24) and secondly from FSNA (code code 3) (Fig. 34 and Fig. 64). The urban development mainly affected the south coastal regions, especially in the area between Portimão and Faro, and was particularly strong nearby to the main metropolitan communities of the northwest and littoral center, namely Porto and Lisbon (Fig. 34 and Fig. 1). Silva and Clarke (2002) described the characteristics and the recent intense urban growth of the metropolitan area of Porto (MAP) and Lisbon (MAL) associated with the economic growth in the end of the XX century (Fernandez-Villaverde et al., 2013). More in details, MAL urban pattern is characterized by a mixture of urban surfaces and large farmlands, an intense urbanization along with train lines and main roads, and the emergence of tertiary centers characterized MAL urban pattern. On the other hand, MAP is described by scattered
urbanization and dispersed settlements, towns and rural villages surrounded by mountains, within small patches of intensive agriculture and pine forests in a steep slope topography. The decline of the rental market in the country at country level lead to the degradation of old urban areas and to the increase of new constructions in the immediate periphery of Lisbon, while in the north new houses were built by the owners in their small plots of land, promoting a more dispersed urban pattern and an irregular spatial growth (Silva and Clarke, 2002). The dispersion of the population and of its activities in MAP is can also be explained and reinforced by the absence of as a consequence of the regional territory planning and the adoption of polycentric models of urban growth adopted by the national authorities (Cardoso, 1996; Silva and Clarke, 2002).

Another active process identified by the performed change analysis is the abandonment of agricultural lands nearby the inland urban areas, which leads to an increase of uncultivated semi-natural and forest areas (Fig. 34) causing an increase of the urban/rural interface.

As regards the RUI definition and mapping model developed in the present study, we tested different buffer width from 100 m to 2-2 km were tested, which led to different RUI's size but resulting in areas of different size, in essence, to approximately equivalent results relatively to the RUI's dynamic. In literature, Vilar et al., (2016) applied a buffer width of 100 m, corresponding to the median of the distances defined in each country’s national legislation (Portugal, Spain, South-France and Italy) for protection against wildfires, which makes brush clearing obligatory within a certain radius around each house located close to forests or scrublands. In US, interface WUI was defined as developed areas in the vicinity of wildland vegetation and mapped considering census blocks above 6.17 housing units/km² that are within a distance of 2.44 km from wildland vegetation (Radeloff et al., 2005).

Finally, we decided to show results obtained applying a buffer width of 1.1 km because smaller values, even if more in line with the Portuguese national indications, could bias the results, given that the spatial resolution of the CLC inventory was of 500 by 500 meters.

RUI definition aimed to map the developed area located in close proximity of wild vegetation, where wildfires can cause deaths, injured, damages to human structures, and finally where human-caused wildfires are more likely to occur. Nevertheless, RUI map was not based on fire incidence measures, thus it not aimed to assess fire risk or fire regimes, which depend on other factors such as topography, climate, weather, vegetation characteristics (Radeloff et al., 2005; Parente and Pereira, 2016). Most of RUI’s area detected in Portugal (Fig. 89) was located in regions of high population density and surrounding major cities, while RUI’s growths mainly occurred in the transition zones from vegetated lands (AA and FSNA) to AS (Fig. 34). The urbanization process and the consequent reconfiguration of Portuguese cities caused new urban problems and challenges associated with the increased fragmentation of the cities and different complex rural-urban relationships, as also reported for Portugal and Spain (Trigal, 2010).

It is important to underline that the impressive increase of the RUI and of BA within the RUI, detected in just a little more than two decades, is not exclusive of Portugal. In Continental US, WUI increased 52% from 1970 to 2000 and 90% of this area included high and highly variable severity fire regimes (Theobald and Romme, 2007). In Europe, Fox et al., (2015) found a progressive increase in fire risk in French Maritime Alps in the period 1960–2009. Badia et al., (2011) noticed that two representative Mediterranean WUI areas in Catalonia were more prone to wildfires in the most recent decade of 2000 than in the 1990s. Pellizzaro et al., (2012) analyzed WUI’s dynamics and landscape changes in a tourist area of North-East Sardinia (Italy) from 1954 to 2008 and discovered that– LULCC was largely associated to a transition from an agro-pastoral economy to one based on tourism. Moreover, they showed an increase in the number of houses and dwellers, which tripled during the study period.
as well as a sharp growth in the length of road network length and, finally, the increase of the WUI’s extension.

The inspection of the accurate Soil Use and Occupancy Chart national map (COS2007-COS2007_v2.0) allowed us to identify the vegetation types and major habitats/plant communities corresponding to each one of the CLC classes for Portugal. Table 3 of the annex shows the composition of the CLC classes in terms of COS classes. For simplicity, results are only presented for COS classes with more than 1% of total CLC area. Nevertheless, it is important to underline that these 19 COS classes account for 98.3% of total CLC area. Despite some expected differences between classifications, essentially noted in CLC classes less affected by wildfires, results obtained for all the other classes can be summarized as follows: (i) Temporary dryland crops is the larger COS class (22% of total CLC area) and accounts for higher area fraction in CLC Agricultural area (code -2) and Scrub and/or herbaceous vegetation associations (code -32); this is particularly consistent with the agricultural practices, especially in southern Portugal; (ii) COS class of Temporary irrigated crops is well distributed over the CLC classes of Rice fields, Permanently irrigated land, Pastures, as well as Beaches, dunes, sands; (iii) COS Shrub classes are particularly present in CLC classes of Shrubs, Pastures as well as in Open spaces with little or no vegetation; (iv) Tree vegetation in COS is also well related to the correspondent CLC classes and allows us to better understand the composition of Forests and Agricultural areas; for example, pure or mixt forests of Cork and Holm oak trees are particularly evident in the CLC classes of Agro-forestry areas, Broad-leaved forests, and Non-irrigated arable lands; on the other hand, Eucalyptus (pure or mixed forests) in COS are important components of the CLC classes of Mixed forests, Complex cultivation patterns, Annual crops associated with permanent crops; and, Broad-leaved forests; finally, pure pinus pinaster forests in COS are comprised in the CLC classes of Coniferous forests, Transitional woodland-shrub, and are especially important in Beaches, dunes, sands, where account for 30% of total area; this finding is in good agreement with the presence of pinus forests in the entire central western coast.

It is important to underline that, from 1990 to 2012, the increase in the BA within the RUI is higher (100%) than the increase in the RUI area (70%). This result suggests that other factors, besides the increase of the RUI area, are responsible of the increase of the burnable area within the RUI. In this regard, it is important to take into account some of the specific characteristics of the country, well described in terms of demographic, territory and forest statistics compiled in Feliciano et al., (2015), which can help to understand the most important factors affecting the forest management in Portugal. Forest is nowadays the dominant land cover in the country (with more than 35% of total area), followed by bushes and grasslands (>29%), and agricultural areas (>24%) (FAO, 2018; Feliciano et al., 2015). According to the National Forest Inventory (IFN, 2010), in the 1995-2010 period, four tree species occupied about 85% of total forest area: Eucalyptus (22%-27%), Cork oak (23%-24%), Maritime pine (30%-23%) and, Holm oak (10%-11%). The first three species have the ability to generate land and business income exceeding 50 euro/ha/year (CM, 2015). Most of the forests and wooded lands (>93% in 1995) have non-industrial private owners and there is a high fragmentation of the forest property, particularly evident in the private sector (Mendes, 2004). Management practices are also very different and changed significantly in the last years, especially in non-industrial private forest (Canadas and Novais, 2014). According to Feliciano et al., (2015), 1/3 of Eucalyptus area is well managed by the industrial pulp and paper companies, with their own forest management and wildfire prevention/fighting resources, while the remaining area is managed by non-industrial private owners, characterized by different objectives and economic logics. In addition, there is a significant heterogeneity in the spatial distribution of all these characteristics (Baptista and Santos, 2005). Small forest holdings (<10 ha), mainly composed by pine and eucalyptus with low profitability, are much more frequent in the northern and central Portugal, while large properties (>100 ha), essentially of cork oak or a complex and unique agroforestry system of cork oak savanna ("montado"), are
predominant in the southern regions of the country. Table 4 of the annex material provides a general description of the main characteristics of forest holdings and forest owners and summarizes the interrelationship between these factors.

Other aspects related to LULCC, such as climate change and biodiversity, are somewhat outside the scope of the present study. However, the abandonment of rural and forest areas and traditional agricultural practices, and the lack of forestry management practices lead to an increase of biomass and fire risk, which can be empowered in a warmer and drier future climate and may have profound impact on ecosystems and biodiversity. For example, the montado, which is composed by sparse cork oak trees and a diversity of understory vegetation (e.g., shrub formations, grasslands), supports higher levels of biodiversity. The decrease in the demand and price of cork has led to a reduction in management practices and to the abandonment of these lands, leading to the invasion of shrubs, reducing the biodiversity and degrading the services provided by these ecosystems (Bugalho et al., 2011).

Results from our analyses shows that in the second half of the investigated period (2000–2006 and 2006–2012) the growth rate of RUI was lower than in the first decade (1990–2000), probably due to a lower growth rate in the process of urbanization of rural areas. Moreover, the decrease of relative BA within the RUI from 2006 to 2012 could be associated with the relative decrease of BA in the last investigated period, as a consequence of recent plans for territorial spatial planning and protection of forest against forest fires (Mateus and Fernandes, 2014; Parente et al., 2016). At European level, urban sprawl’s complexity and magnitude motivated the European Commission to recommend actions and to coordinate land use policies, within the European Cohesion Policy 2007-2013 period (CEC, 2006; EEA, 2006). In Portugal, these efforts were complemented with national programs and regional plans such as the National Policy and Territorial Management (Programa Nacional da Política de Ordenamento do Território, PNPOt) and the Regional Plan for Territorial Planning (Plano Regional de Ordenamento de Território, PROT), supporting the sustainable development and the environmental landscape quality of NUTS-III areas (Noronha Vaz et al., 2012).

However, as far as we know, there is no a specific or general Portuguese legislation about WUI or RUI. It only exists one general mention about RUI in the National Plan to Protect the Forests against Wildfires (CM, 2009). As a measure to protect the urban–forest interface, in this Plan, it is suggested that to protect urban–forest interface is necessary to create and maintain an external buffer strips (10–100 m) around population clusters, especially in those with the highest fire vulnerability, as well as around parks, industrial polygons, landfills, housing, shipyards, warehouses, and other buildings.

Finally, we firmly believe that the results of the present study are sufficiently motivating to promote the development of specific policies and legislation, as well as changes in forest and fire management. The detected increase in growth of the RUI area in Portugal from 1990 to 2012, and particularly of the BA within the RUI, clearly suggests the need of improving fire prevention measures and preparedness policies for this interface region. In fact, as indicated by the Portuguese National Fire Plan 2006 (Oliveira, 2005), the increase of the urban/rural urban interface, as a consequence of the above-mentioned LULCC, causes this area to be under the use of people not educated for fire and unaware of possible source of ignition. In particular, forestry/forest managers must prioritize sustainable forest management practices and make brush clearing obligatory. These paradigm shifts make even more sense if one takes into account that the fire risk is likely to increase in a future climate associated with a higher frequency of longer and more intense extreme events, such as drought and heat waves. Indeed, results of several recent studies suggest a statistical significant increase of the burnt area, not only in Portugal (Pereira et al., 2013) but in the entire Iberian Peninsula (Sousa et al., 2015), for different scenarios of future climate and LULCC (Amato et al., 2018). These findings are in line with the fact that Portugal is the only European country were burnt area did not...
decrease in the last decades (Turco et al., 2016) and with the increase of the future fire danger for the Mediterranean basin countries of Europe and north Africa based on a multi-model ensemble of state-of-the-art regional climate projections (Bedia et al., 2014). In addition, burnt area in Portugal and in the Mediterranean Basin is clearly associated to extreme weather and climate variability, namely the occurrence of heat waves and drought (Amraoui et al., 2015; Pereira et al., 2005; Telesca and Pereira, 2010; Trigo et al., 2006, 2016). Hertig and Tramblay, (2017) results under a future climate scenario suggests an increase in the drought severity and occurrence for the whole Mediterranean region. Russo et al., (2017) have showed that drought promotes a synchronous influence on burned areas over a great part of the Iberian Peninsula the relationship between wildfires and drought events is explained by the influence of the amount of precipitation in winter and spring before summer and by the amplitude and precipitation during summer over almost Portuguese territory. Hertig and Tramblay, (2017) results under a future climate scenario suggests an increase in the drought severity and occurrence for the whole Mediterranean region. Parente et al., (2018) found that all the largest extreme-wildfires in Portugal occurred during or immediately after a heat wave and, a clear association of heat wave characteristics with spatial distribution of the extreme wildfires. Using results of an ensemble of climate models, their findings also suggest: the increase of in the number, duration and amplitude of the heat waves in Portugal for a future climate scenarios (RCP4.5 and RCP8.5), more significant at the end of the 21st century, and for a clearly association of heat wave characteristics with spatial distribution of extreme wildfires. An increase of drought severity in the whole Mediterranean basin is expected in the future. The results of (Hertig and Tramblay, (2017), suggest the increase of drought severity and occurrence for the whole Mediterranean region for future climate scenarios, suggesting All these projection suggest an increase of the fire risk for future climate and burned areas, which should be added to the will impacts of LCLU-the LULCC and to population trends.

6 Conclusions

Continental Portugal registered important land cover changes (about 9% of the whole area) and an increase of the rural-urban interface in the investigated period (1990 – 2012). Most significant changes were associated to transitions from the following CORINE Land Cover classes: Agricultural areas (35.1%) and Forest and semi-natural areas (15.2%) to Artificial surfaces (including Urban areas) and; Agricultural areas to Forest and semi-natural areas (7.3%) and vice-versa (6.3%). However, relative net changes are appreciable only for Artificial surfaces, which registered a substantial increase of about 50%, while Forest and semi-natural areas remained almost constant (0.3%) and Agricultural areas slightly decreased (-4.4%). The spatial distribution of these changes was far from uniform within the territory. Urban sprawl was concentrated in the metropolitan areas of Lisbon and Porto, as well as in central-north and south coastal areas (region of Algarve). A promoted socioeconomic development within the country, the intense rural abandonment and the development of mass tourist industry could act as main drivers for the expansion and reconfiguration of urban areas. On the other hand, in the south and interior north regions we observed assisted to a transition to vegetated land use/land cover types, probably caused by deforestation/afforestation processes and the rural abandonment. The CLC classes mainly affected by these changes where Scrub and/or herbaceous vegetation associations, Forests and Heterogeneous agricultural areas: the increase in Artificial surfaces was precisely due to transitions from these type of land cover. The Vegetated classes with higher burnt area within the RUI detected in the study period were the following: Transitional woodland-shrub, the three types of Forests considered in the CLC inventories and, the three sub-levels of Heterogeneous agricultural areas. These findings suggest the needs of extending the notion concept of wildland-urban interface for
Portugal to rural-urban interface, defined as Forest semi-natural plus Heterogeneous agricultural areas adjacent to Artificial surfaces.

Results of the performed analyses of RUI’s size, burnt areas and burnt areas within the RUI in the four investigated periods (1990, 2000, 2006, and 2012) allow to conclude that, from 1990 to 2012, RUI increased about 70%, burnt area decreased 35% but, nonetheless, burnt area within the RUI increased 100%. These findings underline the need of frequent monitoring and assessment of land use changes and RUI evolution in Portugal, and reinforce the need to focus the attention of forest and fire managers on this highly fire prone region.

The conclusions of this study suggest and encourage more accurate analyses to characterize and mapping the RUI, using accurate and high-resolution data (e.g., real footprints of buildings, road network and census data) to support policymaking and provide practical guidelines for land managers and fire managers. Nevertheless, our study provides precious indications as for what is the global distribution and evolution of RUI in Portugal, identifying which regions need to be prioritized in term of RUI monitoring.

Acknowledgements

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References


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Annexes/supplementary material
Table 3 – Fraction of COS2007v2.0 classes’ area within CORINE land cover classes.

<table>
<thead>
<tr>
<th>Area</th>
<th>&lt; 1 ha</th>
<th>&lt; 5 ha</th>
<th>5 – 20 ha</th>
<th>5 – 100 ha</th>
<th>&gt; 20 ha</th>
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</thead>
<tbody>
<tr>
<td>Forest owners (%)</td>
<td>31%</td>
<td>30%</td>
<td>14%</td>
<td>10%</td>
<td>15%</td>
</tr>
<tr>
<td>Area (%)</td>
<td>10%</td>
<td>16%</td>
<td>12%</td>
<td>7%</td>
<td>55%</td>
</tr>
<tr>
<td>Main tree species</td>
<td>Maritime pine</td>
<td>Maritime pine and chestnut</td>
<td>Eucalyptus</td>
<td>Holm oak and cork oak</td>
<td></td>
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<tr>
<td>Investment</td>
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<td>With investment</td>
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<td></td>
<td></td>
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<tr>
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<td>Management depends on how economy goes</td>
<td>Active management</td>
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<td></td>
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<td>Income</td>
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<td>Forest-enterprise</td>
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<td>Region</td>
<td>Northern and central</td>
<td>Southern</td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Table 4 - Main characteristics of a sample of forest holdings and forest owners studied by Baptista and Santos (2005). Adapted from Feliciano et al. (2015).
Global assessment of land cover changes and rural-urban interface in Portugal related to land cover changes

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Abstract. The wildland-rural-urban interface (WUI/RUI) is a particularly important aspect of known as the fire regime. In Mediterranean basin most of the fires in this pyro region were structures and other human development meet or intermingle with wildland and rural area, is at present a central focus of fire policy and its mapping is crucial for wildfire management and impact. In the Mediterranean basin, the majority of fires are caused by humans and the risk and consequences are particularly high due to the close proximity to population, human of infrastructures and urban of rural/wildland areas. Population increase, RUI’s extension under the pressure of environmental and anthropogenic factors, such as urban growth and the rapid changes in fragmentation of rural areas, deforestation and, more in general, land use occurred in Europe over the last 30 years has been unprecedented, especially nearby the metropolitan areas, and some of these trends are expected to continue. Associated to high socioeconomic development, land cover changes (LULCC) have affected the evolution of RUI in Portugal, from 1990 to 2012, based on LULCC providing also a period with the main objective of investigating how these changes affected RUI’s evolution. Moreover, a qualitative and quantitative characterization of forest fires dynamics in relation to the burnt area within the RUI was performed. Obtained results disclose spatial distribution is far from uniform within the territory. A significant increase in artificial surfaces is registered nearby the main metropolitan communities of the northwest and littoral-central and southern regions, whilst the abandonment of agricultural land nearby the inland urban areas leads to an increase of uncultivated semi-natural and forest areas. Within agricultural areas, heterogeneous patches suffered the greatest changes and were the main contributors to the increase of urban areas. Moreover, these are among the LCLU classes with higher, moreover, this land cover class, together with forests, resulted to be highly affected by wildfires in terms of burnt area, reasons why heterogeneous agricultural areas have been included in the definition of RUI. Finally, the mapped RUI’s area burnt area and burnt area within RUI allow to conclude that from 1990 to 2012 in Portugal, during the investigated period RUI increased in Portugal more than two thirds; nevertheless, burnt area within RUI doubled, which emphasizes the significance of monitoring for land and
This research provides a first quantitative global assessment of RUI in Portugal and presents an innovative analysis on the impact of land use changes on burnt areas.

1 Introduction

Fires can be considered a dynamical ecological factor and an efficient agricultural and landscape management tool but, more recently, they are increasingly considered a hazard (Bond and Keeley, 2005; Fernandes and Botelho, 2003; Hardy, 2005; Moreno and Oechel, 2013; Pyke et al., 2010; Van Wagendonk, 2007). Mediterranean area is particularly affected by forest fires, mainly a consequence of its type of climate and vegetation cover fire proneness (Amraoui et al., 2015; Pellizzaro et al., 2012). In the European countries included in the Mediterranean Basin—Spain, Portugal, Italy and Greece—fire incidence had dramatically increased in the last decades and the average total burnt area has quadrupled since the 60’s (San-Miguel-Ayanz et al., 2012). Portugal stands out from this group of countries for presenting the highest number of forest fires and for being the third most affected country in terms of burnt area in the last three decades (Pereira et al., 2014; San-Miguel-Ayanz et al., 2016). The main sources of ignition in the Mediterranean Europe is associated to human activities (~95%) while only a small percentage of fires...
Forest fires are assuming an increasing importance especially in relation to urban sprawl, which makes it difficult to outline a border between human infrastructures and forest, in order to protect better this vulnerable interface. The modern urban landscape in Europe has a typical star-shaped spatial pattern (Antrop, 2000) where wedge of unchanged countryside can persist between lobes of urban development (Antrop, 2004). The diffuse urban expansion along radiating access roads through industrial zones and dispersed housing, as well as the growth of peri-urban centers in previously rural areas, leave gaps in suburban/exurban space of urban agglomerations (Triegal, 2010). In this context, rural areas nearby and interconnected to large cities and urban areas remain almost unchanged, so that populations and activities, described either as “rural” or “urban”, are closely linked and their distinction is often arbitrary (Tacoli, 1999).

The increase of fire incidence has been considered mainly a consequence of climate and land use changes (Ferreira-Leite et al., 2016; Moreira et al., 2011). Land use changes are at the origin of landscape patterns and dynamics which have a strong influence on forest fires and vice versa (Silva et al., 2011). On the one hand, each vegetated land cover type, such as agricultural, natural and semi-natural vegetation cover, has a specific fire proneness depending on the differences in vegetation structure, moisture content and fuel load composition (Barros and Pereira, 2014; Oliveira et al., 2014; Pereira et al., 2014). Further, fire occurrence affects landscape dynamics by changing the vegetation structure and soil processes according to the fire adaptations of each ecosystem (Pausas et al., 2009; Viedma, 2008).

In Europe, population and urban growth significantly increase during the late XX century, which helps to understand the rapid land cover and land use change (LCLUC) (Noronha Vaz et al., 2012). In European Mediterranean countries LCLUC are mainly characterized by the increasing migration to urban centers at cost of the abandonment of rural areas, and by the expansion of coastal tourism (Alodos et al., 2004; Tedim et al., 2016). The abandonment of low-intensity agricultural lands and grazing practices also caused the increase of forest cover and scrubland vegetation (Millington et al., 2007; Poyatos et al., 2003). Urbanization is defined as the process involving the inhomogeneous transformation of rural and natural landscape into urban and industrial areas, driven by physical conditions and accessibility to the area (Antrop, 2000). The recent urbanization process is characterized by the urban sprawl land use zoning plans, founded on new peri-urban settlements, usually promoted by neighboring municipalities of the central metropolitan city, which is occurring at an unprecedented rate, far beyond the rate of growth of the urban population (Triegal, 2010). The dynamic conversion between rural and urban spaces can become extremely complex (Strubelt and Deutschland, 2001). Urbanization generates the centralization of certain areas by changing the land use, population density, economical activities and transportation network. Expansion and reconfiguration of urban and metropolitan areas comprise the processes of suburbanisation and periurbanisation mainly of coastal regions and, in the interior, of agricultural areas (Triegal, 2010). This complex process causes also the abandonment of remote rural areas, with poor accessibility, population ageing and death, allowing the expansion of natural vegetation and forest (Antrop, 2004).

Significant land use changes were observed in Portugal. Pereira et al. (2011) showed that Portugal presents the highest relative changes of land use among the southern European countries between 2000 and 2006, namely a significant increase in artificial surfaces and sclerophyllous vegetation and a decrease in forest area and natural grasslands, due to rural abandonment.
areas is in concomitance with the abandonment of agricultural land in marginal areas and seems to prevail between 1990 and 2000
while, in the more recent investigated period (2000 – 2006), the intensification of agriculture in areas where irrigation was available
is a predominant process.

The area of interface between the wildland and the urban space, known as Wildland-Urban Interface (WUI), was deeply
investigated by researchers and fire managers in the last decades. According with United States Department of Agriculture (USDA),
WUI was defined as the area “where humans and their development meet or intermix with wildland fuels” (Stewart et al., 2002).
Within the WUI fire can spread readily among vegetation fuels and urban structures. It is well known that anthropogenic features,
such as the distance to roads and houses, negatively influence the probability of forest fire occurrence, while the population density
positively affects it (Conedera et al., 2015; Haight et al., 2004; Hammer et al., 2009; Lampin-

10 – Maillet et al., 2010; Radloff et al., 2005; Stewart et al., 2007), so that the spatial extension of the WUI is determined by these
factors. Urbanization and the consequent abandonment of rural areas caused the expansion of this interface, increasing the
probability that forest fires affect houses and infrastructures (Theobald and Romme, 2007; Zhang et al., 2008). Researchers
developed several geospatial methods for defining and mapping WUI. In Europe, Lampin-Maillet et al. (2010) proposed an
approach and based on the combination of four types of building configuration and three classes of vegetation structure, tested in
the south of France. Following this model, Bouillon et al. (2012) developed WULmap, a software tool for mapping WUI in the
Mediterranean region successfully applied in south-eastern France, eastern Spain and Sardinia in Italy. In the Alpine context,
geospatial approach to map the WUI were developed in Switzerland by Conedera et al. (2015) and in France by Fox et al. (2015).
Pellizzaro et al. (2012) characterized and mapped WUI in Sardinia, Italy, using temporal steps of about 10 years from 1954 and
2008. Other studies focus on additional aspects related to the WUI, namely hazard risk, vulnerability and fire risk
management in Spain (Badia et al., 2011; Galana Martin et al., 2011; Herrero Corral et al., 2012), fuel and fire modelling in
France (Cohen et al., 2003; Pugnet et al., 2013). Most of these studies are local, performed at house-spatial scale or for small
regions within each country.

The WUI concept has to be redefined especially in the European context, taking into account the rural-urban process and associated
landscape changes. Recent studies identified the Rural-Urban Interface (RUI) as the most fire prone area in

15 Mediterranean countries (Badia-Perpinyà and Pallares-Barbera, 2006; Catty et al., 2010; Moreira et al., 2009). In the present
study, RUI is considered and defined for Portugal: the main objective is to investigate RUI’s spatio-temporal evolution at large
scale from 1990 up to 2012. Secondly, it provides a quantitative characterization of forest fires occurred here in relation to the
burnt areas and land covers’ evolution. This research provides a first quantitative global assessment of RUI in Portugal and presents
an innovative analysis on the impact of LCLUC on burnt areas.

20 Mediterranean area is particularly affected by wildfires, mainly as consequence of its type of climate and vegetation cover fire
proneness (Pellizzaro et al., 2012; Angrasai et al., 2015). In the European Mediterranean countries, fire incidence has dramatically increased
in the last decades and the average burnt area (hereafter, BA) has quadrupled since the 60’s (San-Miguel-Ayanz et al., 2012), mainly
due to changes in climate and land use (Vila et al., 2011; Ferreira-Leite et al., 2016). Portugal stands out from this group of countries
since it counts the highest number of wildfires and has been the third most affected country in terms of BA in the last three decades (Pereira
et al., 2014; San-Miguel-Ayanz et al., 2016). On average, about 95% of wildfires with known causes in Europe during the period 1995 to
2010 (corresponding to about 70% of the total number of recorded events) were associated to human activities (Ganteaume et al., 2013),
while only a small percentage of fires (e.g. 1% in Portugal, 5% in Spain) were naturally caused by lightnings (Mateus and Fernandes, 2014;
Vilar et al., 2016). Wildfires have long been considered a dynamic ecological factor and an efficient agricultural and landscape management
tool, but more recently they are increasingly considered a hazard (Fernandes and Botelho, 2003; Bond and Keeley, 2005; Hardy, 2005;
Van Wagtendonk, 2007; Pyke et al., 2010; Moreno and Oechel, 2012), which has motivated governments to implement measures for fire prevention, monitoring and mapping. A key factor in this regards is to study the spatial and temporal distribution of wildfires into a given area, which, being influenced by the underlying socio-economic and environmental factors, is far to be homogeneously distributed. For example a clustered pattern was discovered for Portugal, with fires' hot spots concentrated in the north-west and center regions, while the southern area presented lower densities of wildfires (Pereira et al., 2015; Tonini et al., 2017). Urban sprawl also affect the spatio-temporal pattern of these hazardous events and makes it difficult to define a boundary between human developments and rural areas, in order to better protect this interface where wildfires are more likely to occur. The modern urban landscape in Europe has a typical star-shaped spatial pattern (Antrop, 2000) with wedge of unchanged countryside persisting between lobes of urban development (Antrop, 2004). In this context populations and activities described either as “rural” or “urban” are closely linked and their distinction is often arbitrary (Tacoli, 1998). In the Iberian Peninsula, the diffuse urban expansion along radiating access roads connecting the center to commercial/industrial zones and isolated housing, as well as the growth of peri-urban centers in previously rural areas, leaves gaps in suburban/exurban space of urban populations (Trigal, 2010). Here the expansion and reconfiguration of urban and metropolitan areas comprise the processes of sub-urbanisation and peri-urbanisation, mainly of coastal regions and, in the interior, of agricultural areas (Trigal, 2010). Land use changes are origin of landscape patterns and dynamics and have a strong influence on forest fires (Silva et al., 2011). On the one hand, each vegetated land cover type, such as agricultural, natural and semi-natural vegetation cover, has a specific fire proneness depending on the differences in vegetation structure, moisture content and fuel composition (Barros and Pereira, 2014; Oliveira et al., 2014; Pereira et al., 2014). Further, fire occurrence affects landscape dynamics by changing the vegetation structure and the soil processes according to the fire adaptation of each ecosystem (Viedma, 2008; Pausas et al., 2009).

Population and urban areas significantly increased in Europe during the late XX century, which helps to understand the rapid land cover and use changes (LULCC) (Noronha Vaz et al., 2012). In the European Mediterranean countries, LULCC are mainly caused by the increasing migration to urban centers at cost of the abandonment of rural areas, and by the expansion of costal tourism (Alodos et al., 2004; Tedim et al., 2016). One consequence is the urbanization, defined as the process involving the transformation of rural and natural landscape into urban and industrial areas, caused by the interaction of very different factors and largely influenced by communication, accessibility and mobility (Antrop, 2000). The dynamic conversion among rural and urban spaces can become extremely complex (Strubelt and Deutschland, 2001): urbanization generates the centralization of certain area by changing the land use, population density, economical activities and transportation network. This complex process is at the origin of the abandonment of remote rural areas with poor accessibility, which allows the expansion of wild low vegetation and forest (Antrop, 2004). Specifically, the abandonment of low-intensity agricultural lands and grazing practices caused the increase of forest cover and scrubland vegetation (Povatos et al., 2003; Millington et al., 2007).

Significant LULCC occurred in Portugal in the recent period. The urbanization of coastal areas in the country occurred in concomitance with the abandonment of agricultural land in marginal areas and seemed to prevail between 1990 and 2000 while, in the later period (2000-2006), the intensification of agriculture in areas where irrigation was available was a predominant process (Van Doorn and Bakker, 2007; Diogo and Koomen, 2012). Pereira et al. (2014) observed that among the southern European countries, in the period 2000-2006 Portugal registered the highest rates of land use change marked by a significant increase in artificial surfaces and sclerophyllous vegetation and a decrease in forest area and natural grasslands, because of rural abandonment, urbanization and wildfires.

The interface between the wildland and the urban space, called Wildland-Urban Interface (WUI), has been deeply investigated by researchers and fire managers in the last decades. The United States Department of Agriculture (USDA), defined the WUI as the area “where humans and their development meet or intermix with wildland fuels” (Stewart et al., 2007); here fires can spread readily among vegetation fuels and urban structures. Anthropogenic features, such as the distance to roads and houses, negatively influence the probability of forest fire occurrence, while the population density positively affects it (Haight et al., 2004; Radaloff et al., 2005; Stewart et al., 2007; Hammer et al., 2009; Lampin-Maillet et al., 2010; Conedera et al., 2015), and factors are broadly considered to elaborate WUI maps.
Urbanization and the consequent abandonment of rural areas caused the expansion of this interface, increasing the probability that wildfires affect houses and infrastructures (Theobald and Romme, 2007; Zhang et al., 2008). There are strong evidences that the expansion of the urban and WUI area increased the fire density and related risk (Lampin-Maillet et al., 2011; Fox et al., 2015; Gallardo et al., 2015; Viedma et al., 2015), the cost of houses protection from fire (Pellizzaro et al., 2012) and have an impact on biodiversity and ecosystems (Radeloff et al., 2005). Researchers developed several geospatial models for defining and mapping the WUI, taking into account all the above-mentioned factors. In Europe, Lampin-Maillet et al. (2010) proposed an approach tested in southern France and based on the combination of four types of buildings configuration and three classes of vegetation structure. Following this model, Bouillon et al. (2012) developed WUImap, a software tool for mapping WUI in the Mediterranean region successfully applied in southeastern France, eastern Spain and Sardinia in Italy. In the Alpine context, geospatial approaches to map the WUI were developed in Switzerland by Conedera et al. (2015) and in France by Fox et al. (2015). Pellizzaro et al. (2012) characterized and mapped the WUI in Sardinia, Italy, using temporal steps of about 10 years from 1954 and 2008, and found an increase of the WUI’s extension. Other studies focused on additional aspects related to the WUI, namely hazard/risk, vulnerability and fire risk management in Spain (Badia et al., 2011; Galiana-Martin et al., 2011; Herrero-Coral et al., 2012), fuel and fire modelling in France (Cohen et al., 2003; Pugnet et al., 2013). The majority of these researches were developed locally and performed at house-spatial-scale or for small regions within each country.

The active rural-urban conversion processes and the associated landscape changes, largely documented in the European context, induced to reconsider the WUI concept. In this respect, recent studies defined the Rural-Urban Interface (RUI) as an alternative to the WUI, to highlight the importance of including the rural areas, and identified the RUI as the most fire prone areas in Mediterranean countries (Badia-Perrinà and Pallares-Barbera, 2006; Catry et al., 2010; Moreira et al., 2009). In the present study, authors investigated the RUI in Portugal, the main objective was to analyze changes in land use/land cover occurred in this country in the period 1990-2012 and to assess their impact on the RUI evolution. Moreover, a qualitative and quantitative characterization of the burnt areas within the RUI in relation to the LULCC was performed. Finally, this research provides a first attempt to map the RUI’s extension and evolution at national level for the entire continental Portugal.

2 Study area

Continental Portugal (c.a. of 90,000 km²) is located in the southwest of Iberian Peninsula, bathed by the Atlantic Ocean on the south and west coast and bounded by Spain at north and east. The spatial distribution of the population (was about 10.6 million inhabitants) presents in 2011 and decreased to 10.3 million in 2017; its distribution displays a much higher density in the north-western and southern coastal areas as well around the major cities (Fig. 1). Tagus river divides the country into two regions with approximately similar area but very different in terms of several biophysical and human drivers (Fig. 2) (Parente et al., 2016; Pereira et al., 2015). The north region is characterized by a temperate climate with dry and warm summer (Kottek et al., 2006; Peel et al., 2007), altitude ranging from sea level to about 2000 m and mean watercourse density of about 0.65 km/km². According to CORINE Land Cover (hereafter, CLC) inventory 2012 (EEA, 2016), 54% of north’s area is covered by forests and scrublands, 40% is used for agriculture and about 5% is occupied by artificial surfaces (Fig. 2). The region south of Tagus river is characterized by temperate climate with dry and hot summer (Kottek et al., 2006; Peel et al., 2007), low altitude range (between sea level and about 1000 m), and mean watercourse density of about 0.58 km/km². According to CLC 2012 inventory, 56% of this area is occupied by agricultural areas, 40% by forest and semi-natural areas, and only 2% by artificial surfaces (Fig. 2).
Figure 1 - Population density at parish level (INE, 2012) in the mainland Portugal, with the location of the major cities. NUTS refers to regions according with Nomenclature of Territorial Units for Statistics level II (Eurostat, 2015; Santos, 2014).

Figure 2 – Land cover of mainland Portugal based on the second level of CORINE land cover inventory 2012 (EEA, 2016)
Figure 1 - Population density at parish level (INE, 2012) in the mainland Portugal, with the location of the major cities. NUTS refers to regions according to Nomenclature of Territorial Units for Statistics level II (Eurostat, 2017; Santos, 2014).
3 Data and methodology

Forest fires (Wildfires) digital data came from the National Mapped Burned Area dataset (NMBA), provided by the Portuguese Institute of Conservation of Nature and Forests (ICNF) for the period 1990-2015. The official NMBA was complemented with data from 1975 to 1989, provided by the research team from the Instituto Superior de Agronomia (ISA) which...
produced the official database (Barros and Pereira, 2014; Oliveira et al., 2012). The final official National fire database covers the 1975-2015 period and comprises about 49,000 fire events for a total of burnt area (BA) of 4,430,000 ha. (Oliveira et al., 2012; Barros and Pereira, 2014). The annual fire perimeter maps resulted from semi-automatic supervised image classification (Gorte, 1999) performed with the classification and regression trees algorithm of Breiman et al. (1984) on late summer-autumn Thematic Mapper/Enhanced Landsat satellite imagery. Technological improvements of satellite sensors allowed acquiring and processing over time higher-resolution images with an increasing accuracy from the initial 30 hectares to 5 hectares after 1984, and even higher after 2005. To ensure the accuracy of the data, the resulting fire perimeters maps were compared against field statistics gathered on the ground by the National Forest Authority, and by the National Civil Protection Authority (Barros and Pereira, 2014; Oliveira et al., 2012). For each fire record, the dataset comprises the burnt area (BA, perimeter map) and the year of occurrence. Thanks to technological improvements, higher resolution original images were acquired and processed over time allowing to increase the accuracy to 5 hectares after 1984 and even higher after 2005. For homogeneity reasons, the present study was performed using forest fires with the minimum area resolution fixed for consistency at 5 ha.

Land use/land cover’s information comes from the CORINE Land Cover CLC inventory (CLC) provided by European Environment Agency (EEA). CLC is delivered as cartographic product, both in raster and vector-shapefile (i.e., a regular grid of cells) and in vector-shapefile (as point, line and polygons) format. The minimum cartographic unit is of 25 ha (500 by 500 m) with a geometric accuracy of 100 m minimum and a thematic accuracy over 85 percent (EEA, 1994). CLC nomenclature comprises 44 land cover classes grouped in a three-level hierarchy. This hierarchical classification system with 44 classes at the third and most detailed level (Table 1). The five more general classes for the first level arc: Artificial Surfaces (AS), Agricultural Areas (AA), Forest and Semi-Natural Areas (FSNA), Wetlands, Water bodies. CLC inventories are currently available for four periods (1990, 2000, 2006, 2012) with a minimum time consistency of plus/minus one year. CLC has already been used for land-use change and urban dynamics studies, namely in Portugal (Noronha Vaz et al., 2012). This study used (Noronha Vaz et al., 2012; Pereira et al., 2014) to identify and detail the major habitats/plant communities/vegetation types corresponding to each CLC class in Portugal, we employed the Soil Use and Occupancy Chart (Carta de Uso e Ocupação do Solo, COS) provided by the Portuguese Directorate-General of the Territory (Direção-Geral do Território, DGT). DGT is the national public body responsible for pursuing public policies for land use and town planning. We compared CLC2006 with COS2007v2.0 because these are the closest inventories (in time) within the study period. In addition, COS2007v2.0 presents improvement from the thematic and geometric point of view (DGT, 2016; Sarmento et al., 2016): it includes 225 classes (32 more than the initial version) at the most detailed level, distributed over 5 hierarchical levels. In the present study, the four CLC inventories were employed to analyze LULCC at different levels and to map RUI at different periods, according to the methodology described below.

The first analysis consisted in investigating LULCC within the LCLUC between either end of the study period (i.e., from 1990 to 2012) which allowed to elaborate the map of changes showing the transitions among all the classes. A detailed analysis was carried out on LCLUC to explore the five more general classes (CLC first level hierarchy) and, more in detail, to quantify gained and lost areas with reference to both the first and the second level hierarchy of CLC. Moreover, the difference between gains and losses within different land cover classes, with particular regard to the classes most representative of RUI, then, spatial and temporal distribution of BA...
On each land cover, each class divided by the total area covered by the specific class in the later period (i.e. the net change) was computed and analysed to highlight which class is most affected by forest fires in each investigated period. To this end, we considered the cumulative burnt surface within the three-years’ time consistency for CLC (i.e. CLC_year ± 1 year). Finally, RUI was mapped using a geospatial approach (Fig. 3) the result expressed in percentage.

25—designed to extract the area of intersection between a buffer of 1 km around the Artificial Surfaces (AS) and the area resulting from the sum of the Forest and Semi-Natural Area (FSNA) plus the Heterogeneous Agricultural Areas (HAA). Other agricultural areas were not included in the RUI since arable lands and permanents crops are usually well managed, mostly irrigated and frequently constitute an obstruction to fire spread which, for these reasons, present much lower fire incidence in comparison to HAA and other vegetated CLC classes. The geocomputation was implemented into a Model Builder in ArcGIS™ environment and performed over the four periods, corresponding the available versions of CLC.

RUI was then mapped for each period (1990, 2000, 2006 and 2012) using a geospatial approach designed to extract the area of intersection between a buffer around the AS and the area resulting from the sum of the FSNA plus the Heterogeneous Agricultural Areas (HAA), a sub-level of AA). Different buffers from 100 m to 2 km were tested: finally, we adopted a buffer width of 1 km, corresponding to two times the spatial resolution of CLC inventories. This value is in line with values applied in other countries for WUI mapping (Radeloff et al., 2005; Vilar et al., 2016) and, in the same time, it is a large enough bias in the results. The agricultural areas (i.e. arable lands, permanents crops and pastures) were not included in the RUI definition since these vegetated land covers are usually well managed, mostly irrigated and frequently constitute an obstruction to fire spread. Similarly, San-Miguel-Ayanz et al., (2012) suggested that HAA have to be considered in the definition and quantification of the RUI in Portugal, together with FSNA.

The geocomputation which allowed producing the RUI’s maps was performed under ArcGIS™ software environment. Namely, the geoprocessing workflows were implemented into a Model Builder (Fig. 3), a specific application used to create, edit and manage models, meant as workflows that bundle together sequences of geoprocessing tools (e.g. selection, buffer, intersect), feeding the output of one tool into another tool as input (i.e. the raster or vector digital data). Finally, we analyzed how each land cover class (i.e. level 3 of the third level hierarchy of CLC) was affected by wildfires. The sum of BA for each investigated period within the RUI. To this end, polygons defining the BA registered at each CLC year plus/minus one year (1989-1991, 1999-2001, 2005-2007, 2011-2013) were merged together and the resulting BA polygons were clipped over the corresponding RUI map. The resulting outputs, representing the BA within the RUI cumulated over three years around each investigated period, were finally overlapped with the CLC source map.
Table 1: CORINE Land Cover (CLC) nomenclature. Numbers on the left represent the CLC code. (Source: EEA, 2016)

<table>
<thead>
<tr>
<th>CLC Code</th>
<th>Category</th>
</tr>
</thead>
<tbody>
<tr>
<td>13</td>
<td>Mining, dump and construction sites</td>
</tr>
<tr>
<td>13.1</td>
<td>Mineral extraction sites</td>
</tr>
<tr>
<td>13.2</td>
<td>Bauxite sites</td>
</tr>
<tr>
<td>13.3</td>
<td>Construction sites</td>
</tr>
<tr>
<td>14</td>
<td>Artificial, non-agricultural vegetated areas</td>
</tr>
<tr>
<td>14.1</td>
<td>Forest urban areas</td>
</tr>
<tr>
<td>14.2</td>
<td>Land and feature boundaries</td>
</tr>
<tr>
<td>2</td>
<td>Agricultural areas</td>
</tr>
<tr>
<td>2.1</td>
<td>Arable land</td>
</tr>
<tr>
<td>2.1.1</td>
<td>Non-irrigated arable land</td>
</tr>
<tr>
<td>2.1.2</td>
<td>Permanently irrigated land</td>
</tr>
<tr>
<td>2.1.3</td>
<td>Rice fields</td>
</tr>
<tr>
<td>2.2</td>
<td>Permanent crops</td>
</tr>
<tr>
<td>2.2.1</td>
<td>Vineyards</td>
</tr>
<tr>
<td>2.2.2</td>
<td>Fruit trees and berry plantations</td>
</tr>
<tr>
<td>2.2.3</td>
<td>Olive groves</td>
</tr>
<tr>
<td>2.2.4</td>
<td>Pomegranate plantations</td>
</tr>
<tr>
<td>2.3</td>
<td>Pastures</td>
</tr>
<tr>
<td>2.4</td>
<td>Heterogeneous agricultural areas</td>
</tr>
<tr>
<td>2.4.1</td>
<td>Annual crops associated with permanent crops</td>
</tr>
<tr>
<td>2.4.2</td>
<td>Complex cultivation patterns</td>
</tr>
<tr>
<td>2.4.3</td>
<td>Land principally occupied by agriculture, with significant areas of natural vegetation</td>
</tr>
<tr>
<td>2.4.4</td>
<td>Agro-forestry areas</td>
</tr>
<tr>
<td>3</td>
<td>Forest and semi natural areas</td>
</tr>
<tr>
<td>3.1</td>
<td>Forests</td>
</tr>
<tr>
<td>3.1.1</td>
<td>Broad-leaved forest</td>
</tr>
<tr>
<td>3.1.2</td>
<td>Coniferous forest</td>
</tr>
<tr>
<td>3.1.3</td>
<td>Mixed forest</td>
</tr>
<tr>
<td>3.2</td>
<td>Scrub and/or herbaceous vegetation associations</td>
</tr>
<tr>
<td>3.2.1</td>
<td>Natural grasslands</td>
</tr>
<tr>
<td>3.2.2</td>
<td>Moors and heathland</td>
</tr>
<tr>
<td>3.2.3</td>
<td>Sclerophyllous vegetation</td>
</tr>
<tr>
<td>3.2.4</td>
<td>Transitional woodland-shrub</td>
</tr>
<tr>
<td>3.3</td>
<td>Open spaces with little or no vegetation</td>
</tr>
<tr>
<td>3.3.1</td>
<td>Beaches, dunes, sands</td>
</tr>
<tr>
<td>3.3.2</td>
<td>Bare rocks</td>
</tr>
<tr>
<td>3.3.3</td>
<td>Sparsely vegetated areas</td>
</tr>
<tr>
<td>3.3.4</td>
<td>Burnt areas</td>
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<tr>
<td>3.3.5</td>
<td>Glaciers and perennial snow</td>
</tr>
<tr>
<td>4</td>
<td>Wetlands</td>
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<tr>
<td>4.1</td>
<td>Inland wetlands</td>
</tr>
<tr>
<td>4.1.1</td>
<td>Inland marshes</td>
</tr>
<tr>
<td>4.1.2</td>
<td>Peat bogs</td>
</tr>
<tr>
<td>4.2</td>
<td>Maritime wetlands</td>
</tr>
<tr>
<td>4.2.1</td>
<td>Salt marshes</td>
</tr>
<tr>
<td>4.2.2</td>
<td>Shingles</td>
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<tr>
<td>4.2.3</td>
<td>Estuarine flats</td>
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<tr>
<td>5</td>
<td>Water bodies</td>
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<tr>
<td>5.1</td>
<td>Inland waters</td>
</tr>
<tr>
<td>5.1.1</td>
<td>Water courses</td>
</tr>
<tr>
<td>5.1.2</td>
<td>Water bodies</td>
</tr>
<tr>
<td>5.2</td>
<td>Marine waters</td>
</tr>
<tr>
<td>5.2.1</td>
<td>Coastal lagoons</td>
</tr>
<tr>
<td>5.2.2</td>
<td>Estuaries</td>
</tr>
<tr>
<td>5.2.3</td>
<td>Mud and oyster beds</td>
</tr>
</tbody>
</table>

Figure 3 - Framework implemented in the Model Builder (ArcGIS™) to map the rural-urban interface. CLC=CORINE land cover; AS=artificial surfaces; FSNA=forest and semi natural areas; HAA=Heterogeneous agricultural areas.

4 Results

4.1 LULC change analysis

The LULC transition map (Fig. 4) was elaborated considering analysis of main changes in the area occupied by the first level hierarchy classes between CLC1990 and CLC2012. The results allowed to visualize the main transitions occurred within the entire investigated period (Fig. 4) and thus to have an overview of the LULCC occurred in Portugal. It resulted that
transitions occurred between vegetated areas (i.e., AA and/or FSNA) to artificial surfaces (AS) and the transition between FSNA and AA in both directions. AS increased mainly nearby the main metropolitan communities of the northwest and littoral central and southern regions. An A transition from vegetated areas (AA and FSNA) to AS is also visible in centre-north and is probably due to the intensification of the main road network, which takes the place of agricultural and semi-natural areas, probably to connect the emergent inhabited rural area. The conversion from FSNA to AA and vice-versa appeared to be an active and dynamic process predominating in the inner northern region and especially prevailing in the southern half of the country. Changes in surface were estimated, but it was revealed also in the inner northern region. Figure 5 shows the areas gained and lost for each CLC first-level class in absolute (as area gained and lost) and relative values (and the net percentage of changes, computed relatively to the total area of each class in the later land cover and expressed as percentage) (Fig. 5). The main changes in terms of surface area were registered by AS, which increased 165 × 10^3 ha, and AA, which decreased 184 × 10^3 ha. These values have a completely different impact. In terms of net percentage of change, the net increase of AS is of about 50%, while the net decrease of AA is of 4.4%. The two classes which mainly contributed to the increase in AS were AA, with 110 × 10^3 ha, and FSNA, with 50 × 10^3 ha.
Figure 4 – Map of land cover/land use transition from 1990 and 2012, evaluated considering the first level hierarchy of CLC 1990 and CLC 2012
Figure 5 – Area gained and lost from 1990 to 2012 for each CORINE land cover class, considering the first level hierarchy. Net percentage changes were computed relatively to the total area of each class in the later land cover.

A more detailed analysis was carried out to investigate changes from 1990 to 2012 occurred within classes considering the second level hierarchy. Table 1 shows the majority of the CLC sub-classes displayed important net changes in terms of relative gains and losses compared with values for the same classes in the later period. Scrub and/or herbaceous vegetation associations (code 32) registered a net gain of about $520 \times 10^3$ ha ($+24\%$), while the Forest area decreased (code 31) decreased...
about $460 \times 10^3$ ha (-23%). Arable land (code 21) was the only agricultural areas AA registering an important negative net change of $225 \times 10^3$ ha (-20%). Among artificial surfaces AS, Urban fabric (code 11) significantly increased $110 \times 10^3$ ha (45%), and, in terms of net percentage of change by class, all the other three AS sub-levels, including Industrial/commercial and transport unit (code 12), Mine/dump and construction sites (code 13), Artificial/non-agricultural vegetated areas (code 14), increased more than half.

<table>
<thead>
<tr>
<th>CLC level 1</th>
<th>CLC level 2</th>
</tr>
</thead>
<tbody>
<tr>
<td>1 Artificial surfaces</td>
<td>1.1 Urban fabric</td>
</tr>
<tr>
<td>1.2 Industrial, commercial and transport units</td>
<td></td>
</tr>
<tr>
<td>1.3 Mine, dump and construction sites</td>
<td></td>
</tr>
<tr>
<td>1.4 Artificial, non-agricultural vegetated areas</td>
<td></td>
</tr>
<tr>
<td>2 Agricultural areas</td>
<td>2.1 Arable land</td>
</tr>
<tr>
<td>2.2 Permanent crops</td>
<td></td>
</tr>
<tr>
<td>2.3 Pastures</td>
<td></td>
</tr>
<tr>
<td>2.4 Heterogeneous agricultural areas</td>
<td></td>
</tr>
<tr>
<td>3 Forest and semi-natural areas</td>
<td>3.1 Forests</td>
</tr>
<tr>
<td>3.2 Scrub and/or herbaceous vegetation associations</td>
<td></td>
</tr>
<tr>
<td>3.3 Open spaces with little or no vegetation</td>
<td></td>
</tr>
<tr>
<td>4 Wetlands</td>
<td>4.1 Inland wetlands</td>
</tr>
<tr>
<td>4.2 Maritime wetlands</td>
<td></td>
</tr>
<tr>
<td>5 Water bodies</td>
<td>5.1 Inland waters</td>
</tr>
<tr>
<td>5.2 Marine waters</td>
<td></td>
</tr>
</tbody>
</table>

Table 1 - CORINE land cover (CLC) first and second level hierarchy.
Figure 6 – Area lost and gained from 1990 to 2012 for each CORINE land cover class, considering the second level hierarchy. Net percentage changes were computed relatively to the total area of each class in the later land cover.

The bar graph shows the contributions to net changes in the artificial surfaces sub-classes (Fig. 7) shows that Urban fabric, which includes buildings, roads and artificially surfaced areas, grew at the expense almost exclusive of heterogeneous agricultural areas, while HAA (code 24). On the other hand, the increase of Industrial commercial and transport was mainly due to the decrease of forest, heterogeneous agricultural area, forests (code 31), HAA (code 24) and Scrub and/or herbaceous vegetation associations (code 32).
Figure 7 – Contribution to the net changes from 1990 to 2012 of “Urban fabric” (orange bars) and “Industrial, commercial and transport” and “Urban fabric” (blue bars) from the other CLC sub-classes.

4.2 Spatial distribution and characterization of burned areas

The overlapping between polygons identifying burnt area (BA) and the CORINE Land Cover maps was performed. Almost all the CLC third-level classes belonging to investigate the land cover types most FSNA (code 3) were affected by forest fires. In this detailed analysis, the third level hierarchy of the CLC was considered and values computed for the four available periods (1990, 2000, 2006, and 2012). Therefore, burnt areas were aggregated over the three years around each period (i.e. CLC_year ± 1). Fires affected almost all the sub-classes wildfires in terms of the first level hierarchy forest.
and semi-natural burned area (Table 12), with the Transitional woodland-shrub (code 324) and Mixed forest (code 313) as the first and second more damaged classes in each period. This trend was similar in all the four investigated frame-periods, as highlighted in Fig. 8, which reveals that the sum of Burnt RUI, as the ratio of BA over the total BA for the entire frame period. The peak of BA in Transitional woodland-shrub equal (code 324) equals to about $43 \times 10^3$ ha in the period 2005-2007, compared with about $15 \times 10^3$ ha in 2011-2013, $14 \times 10^3$ ha in 1999-2001 and $10 \times 10^3$ ha in 1989-1991. It also emerges that the three over four sub-classes levels of heterogeneous agricultural areas (HAA) (code 243, 241, 242) are more highly affected by fires than other type wild scrub vegetation. Consequently, this agricultural land cover is more vulnerable to forest fires and has to be included in the wildfires, thus confirming the need of including HAA in the RUI’s definition of the interface zone between urban space and wildland-rural areas.

<table>
<thead>
<tr>
<th>CLC code</th>
<th>CLC classes</th>
<th>BA within RUI (ha)</th>
</tr>
</thead>
<tbody>
<tr>
<td>324</td>
<td>Transitional woodland-shrub</td>
<td>6086.31</td>
</tr>
<tr>
<td>313</td>
<td>Mixed forest</td>
<td>4368.85</td>
</tr>
<tr>
<td>312</td>
<td>Coniferous forest</td>
<td>3104.72</td>
</tr>
<tr>
<td>322</td>
<td>Moors and heathland</td>
<td>1835.03</td>
</tr>
<tr>
<td>243*</td>
<td>Land principally occupied by agriculture, with significant areas of natural vegetation</td>
<td>1535.36</td>
</tr>
<tr>
<td>311</td>
<td>Broad-leaved forest</td>
<td>1144.72</td>
</tr>
<tr>
<td>241*</td>
<td>Annual crops associated with permanent crops</td>
<td>698.75</td>
</tr>
<tr>
<td>242*</td>
<td>Complex cultivation patterns</td>
<td>677.27</td>
</tr>
<tr>
<td>321</td>
<td>Natural grasslands</td>
<td>638.39</td>
</tr>
<tr>
<td>323</td>
<td>Sclerophyllous vegetation</td>
<td>562.68</td>
</tr>
<tr>
<td>334</td>
<td>Burnt areas</td>
<td>396.87</td>
</tr>
<tr>
<td>333</td>
<td>Sparsely vegetated areas</td>
<td>321.85</td>
</tr>
<tr>
<td>332</td>
<td>Bare rocks</td>
<td>123.84</td>
</tr>
<tr>
<td>244*</td>
<td>Agro-forestry areas</td>
<td>17.25</td>
</tr>
<tr>
<td>351</td>
<td>Beaches, dunes, sands</td>
<td>5.22</td>
</tr>
</tbody>
</table>

(*) Heterogeneous agricultural areas

Table 12 - Classes of land use, as defined by the third level hierarchy of the CORINE Land Cover (CLC), affected by forest fires in terms of Burned Area (BA) within the Rural Urban Area (RUI) during three investigated frame periods (1990, 2000, 2006, 2012).
Figure 8 – Classes of land use, as defined by the third level hierarchy of the CORINE Land Cover, affected by forest fires in terms of Burned Area within the Rural Urban Interface during three investigated frame periods (1990, 2000, 2006, 2012), expressed as percentage wildfires expressed in percentage as the ratio of BA affecting each CLC classes over the total BA for each frame period.

4.3 RUI map

Evidence from the previous analyses indicates that heterogeneous agricultural areas (classes FSNA and HAA) have to be...
Included were considered in the computation of the present study to describe the flammable rural area, which, intermingling with the urban interface, the land cover class contributes, with forest and semi-natural area (FSNA), to delimiting, defines the vegetated land area. RUI. Thus, RUI maps finally result from the zone of intersection between the sum of HAA plus FSNA and an enlarged area around artificial surfaces (AS). A result (Fig. 3). The result was a dynamic zone of interface which evolves in space and in time following the LULC changes due to LULCC (Fig. 7). Analyzing the period 1990-2012, the increase of RUI was more active in the north-west and along the coast, where the transition from heterogeneous agricultural areas (HAA) to urban fabric was particularly intense. This evolution was mainly due to the urban growth and to the intensification of the road network.

Figure 9 – Maps of the rural-urban interface (RUI) in Portugal estimated for the different periods of investigation, and corresponding to the available CORINE Land Cover inventories (1990, 2000, 2006, 2012)
Globally and in the investigated periods, some trends must be emphasized. Taking into account that the period between CLC inventories is not constant, RUI increased in time from about 780×10^3 ha in 1990 up to about 131×10^3 ha in 2012 following a power-law (RUI=776310.year^{0.1686}, R^2=0.99). Total BA presents a slight decreasing trend (~6×10^3 ha/year) but, essentially, fluctuates between higher (BA > 450×10^3 ha), in 1990 and 2006, and lower values (BA ~ 310×10^3 ha), in 2000 and 2012. BA within the RUI slightly increase over time (~1.5×10^3 ha/year), however doubled (7% to 15%) between 2000 and 2012 and tripled (4% to 15%) between 1990 and 2006.

The total size of the RUI, the fraction of BA within the RUI (BAR) and the total BA (TBA) were computed (Fig. 10). It resulted that RUI increased from about 780×10^3 ha in 1990 up to about 1310×10^3 ha in 2012 following a power-law (RUI=776310.year^{0.1686}, R^2=0.99). Moreover, we computed the contribution both to the RUI and to the BA within the RUI (BAR) of each CLC class that make up the RUI, namely HAA (code 24), Forests (code 31), Scrub and/or herbaceous vegetation associations (code 32), and Open spaces with little vegetation (code 33). Results can be summarized as follows (Fig. 10): (a) the relative contribution of the those four CLC classes to the RUI increases in time at approximately the same rate; (b) HAA (code 24) is the CLC class with the largest area within the RUI (~50%); (c) in terms of BA within the RUI (BAR), the most affected class is the Scrub and/or herbaceous vegetation associations (code 32), followed by Forest (code 31), HAA (code 24) and then Open spaces with little or no vegetation (code 33). The total extent of the BA (TBA) fluctuated, with a maximum in 1990 (about 500×10^3 ha) followed by 2006 (~460×10^3 ha), while in 2000 and 2012 its value was lower and equal to about 310×10^3 ha. The portion of BA included within the RUI (RBA), expressed as percentage over the total BA, tends to increase in time, passing from 4% and 7% in 1990 and 2000 to 15% and 14% in 2006 and 2012, respectively.
Figure 10 – Evolution of the absolute area of the rural-urban interface (RUI), total burnt area within the RUI (BAR), total burnt area in the RUI (TBA), and relative burnt area in the RUI (BA in RUI), from RBA, as %, for each investigated period (1990 to 2000, 2006, 2012).

Discussion

The LCLU analysis performed in the present study indicates that from 1990 to 2012 artificial surfaces AS (code 1) globally increased in Portugal by about 50% (Fig. 5). This growing process is in good agreement with previous findings of other authors (Tavares et al., 2012; Marques et al., 2014; Meneses et al., 2014; Oliveira et al., 2017). Moreover, the present study confirms that the urban growth is mainly due to land cover process in Portugal (quantified as changes of areas previously in AS) was principally occupied by heterogeneous agriculture and secondly by forest and semi-natural land covers from FSNA (code 3) (Fig. 4 and Fig. 7). This growing process is evident in the south coastal regions, mostly in the area between Portimão and Faro, but is strongest and was particularly strong nearby to the main metropolitan communities of the northwest and littoral centre, namely Porto and Lisbon (Fig. 4 and Fig. 1). These patterns of changes are in very good agreement with previous findings of other authors. According to Marques et al. (2014), housing stock statistics reveal a significant increase in the last decades, namely: 30% of the buildings existing in 2011 were built in the previous two decades, especially in the Algarve (24%), due to tourism (construction of hotels and second residence houses), Centro (12%) and the lowest in Norte and Alentejo (10%).

It is also important to understand the major features and drivers of urban development in these areas as well as main differences between regions, namely the metropolitan areas of Porto and Lisbon (MAP and MAL, respectively). Silva and Clarke (2002) described the characteristics and the recent intense urban growth of the metropolitan area of Porto (MAP) and Lisbon (MAL) associated with the economic growth in the end of the XX century (Fernandez-Villaverde et al., 2013). More in details, MAL urban pattern is characterized by (i) a mixture of urban surfaces and large farmlands (olive, cork, and fruit orchards); (ii) fast urban sprawl from main cities (Lisbon, Oeiras, Cascais, Setúbal, Almada and Barreiro); (iii) an intense urbanization along with train lines and main roads-diffusing from those major urban centres; and (iv) the emergence of tertiary centres and spread of low-density residential areas based on private car transport. On the other hand...
hand, MAP is described by (i) scattered urbanization with high density in major cities (Porto, Matosinhos and Vila Nova de Gaia), (ii) dispersed settlements, towns and rural villages, surrounded by mountains, within small patches of intensive agriculture and pine forests, in a steep slope topography, and (iii) intense, unorganized and irregular urbanization growth from these
scattered urbanization patches. Main drivers of the fast urbanization growth in XX century helps to better understand the LCLUC patterns, and comprises: (i) urban rent control (Malpezzi and Rydell, 1986; Marques et al., 2014); (ii) municipal autonomy law, which allowed municipalities to issue licenses to build new houses and provide urban infrastructure services (Da Cruz and Marques, 2011; Delgado, 2014); (iii) easy access to credit market with low interest and mortgage rates for housing, construction and road infrastructure (Barradas et al., 2011; Fernandez-Villaverde et al., 2013; Marques et al., 2014; Rebeiro, 2007); (iv) liberalization of the real estate market and laws that protect tenants; (v) the return of the Portuguese habitants mainly from the former African colonies after the revolution of April 25, 1974 (Marques et al., 2014). The consequent decline of the rental market and road infrastructure lead to the degradation of old urban areas and the increased constructions in the immediate periphery in the case of Lisbon, while in the north, new houses were built by the owners in their small plots of land, promoting a more dispersed urban pattern and an irregular spatial growth, jeopardizing the viability plan new settlements (Silva and Clarke, 2002). The dispersion of the population and its activities in MAP is also explained and reinforced by the absence of a regional territory planning and the adoption of polycentric models of urban growth by the national authorities (Cardoso, 1996; Silva and Clarke, 2002). Another active process identified by the performed change analysis is the abandonment of agricultural lands nearby the inland urban areas, which leads to an increase of uncultivated semi-natural and forest areas (Fig. 4). The rapid growth of metropolitan areas causes an increase of the urban/rural interface.

As regards the RUI definition and mapping model developed in the present study, we tested different buffer width from 100 m to 2 km, which led to different RUI’s size but, in essence, to approximately equivalent results relatively to the RUI’s dynamic. In literature, Vilar et al. (2016) applied a buffer width of 100 m, corresponding to the median of the distances defined in each country’s national legislation for protection against fires, which makes brush clearing obligatory within a certain radius around each house located close to forests or scrublands. In US, WUI was defined and mapped at a more global scale, considering census blocks above 6.17 housing units/km² within a distance of 2.4 km from wildland vegetation (Radeloff et al., 2005). In the present study, we applied a buffer width of 1 km, which is in between of the 100 m used in southern Europe and the 2.4 km used in US. Different buffer width were also tested, leading to different RUI size but, in essence, to approximately equivalent relative results. Regarding the buffer width applied to delimitate this interface, Vilar et al. (2016) used 100 meters; this value corresponds to the median of the distances defined in each country’s national legislation for protection against fires, which makes brush clearing obligatory within a certain radius around each house located close to forests or scrublands. In US, WUI was defined and mapped at a more global scale, considering census blocks above 6.17 housing units/km² within a distance of 2.4 km from wildland vegetation (Radeloff et al., 2005). In the present study, we applied a buffer width of 1 km, which is in between of the 100 m used in southern Europe and the 2.4 km used in US. Different buffer width were also tested, leading to different RUI size but, in essence, to approximately equivalent relative results. As regards the RUI definition and mapping model developed in the present study, we tested different buffer width from 100 m to 2 km, which led to different RUI’s size but, in essence, to approximately equivalent results relatively to the RUI’s dynamic. In literature, Vilar et al. (2016) applied a buffer width of 100 m, corresponding to the median of the distances defined in each country’s national legislation (Portugal, Spain, South-France and Italy) for protection against wildfires, which makes brush clearing obligatory within a certain radius around each house located close to forests or scrublands. In US, interface WUI was defined as developed areas in the vicinity of wildland vegetation and mapped considering census blocks above 6.17 housing units/km² that are within a distance of 2.4 km from wildland vegetation (Radeloff et al., 2005). Finally, we decided to show results obtained applying a buffer width of 1 km because smaller values, even if more in line with the Portuguese national indications, could bias the results, given that the spatial resolution of the CLC inventory was of 500 by 500 meters.
Most of RUI areas detected in Portugal (Fig. 9) were located in regions of high population density and surrounding major cities, while RUI increase mainly occurred in the transition zones from agricultural areas and vegetated lands (AA and forest and semi-natural areas FSNA) to artificial surfaces (Fig. 4). Urbanization and the consequent reconfiguration of Portuguese cities caused new urban problems and challenges associated to the increased fragmentation of the cities and different rural-urban relationships, affecting small as well as medium and large cities, as also reported for Portugal and Spain (Trigal, 2010). It is important to underline that the impressive increase of the BA within the RUI, detected in just a little more than two decades, is not exclusive of Portugal. In Continental US, WUI increased of 52% from 1970 to 2000 and 90% of this area included high and highly variable severity forest fire regimes (Theobald and Romme, 2007). In Europe, Fox et al. (2013) Fox et al. (2013) found a progressive increase in fire risk in French Maritime Alps during approximately the same period (1960—2009). Badia et al. (2011). Badia et al. (2011) noticed that two representative Mediterranean WUI areas in Catalonia were more prone to forest fires in the most recent decade of 2000 than in the 1990s. Pellizzaro et al. (2012) analyzed WUI’s dynamics and landscape changes in a tourist area of North-East Sardinia (Italy) from 1954 to 2008. This region evolution resembles what happens in Algarve in last decades, reason why their findings are worth noting here and comprise: (i) large LULCC and discovered that LULCC was largely associated to a transition from an agro-pastoral economy to one based on tourism; (ii) gradual and steady increase in the number of houses and dwellers, which tripled during the study period, as well as a sharp growth in summer population load which can be 10 times greater than in other seasons; (iii) rose of road network length; (iv) and finally, increase in the WUI’s presence, extension (particularly sharp in WUI’s length) as well as intensification of the different types of WUI (isolated and scattered houses and clustered buildings).

The inspection of the accurate Soil Use and Occupancy Chart national map (COS2007 v2.0) allowed us to identify the vegetation types and major habitats/plant communities corresponding to each one of the CLC classes for Portugal. Table 3 of the annex/supplementary material, shows the composition of the CLC classes in terms of COS classes. For simplicity, results are only presented for COS classes with more than 1% of total CLC area. Nevertheless, it is important to underline that these 19 COS classes account for 98.3% of total CLC area. Despite some expected differences between classifications, essentially noted in CLC classes less affected by wildfires, results for all the other classes can be summarized as follows: (i) Temporary dryland crops is the larger COS class (22% of total CLC area) and accounts for higher area fraction in CLC Agricultural area (code 2) and Scrub and/or herbaceous vegetation associations (code 32); this is particularly consistent with the agricultural practices, especially in southern Portugal; (ii) COS class of Temporary irrigated crops is well distributed by CLC classes of Rice fields, Permanently irrigated land, Pastures, as well as Beaches, dunes, sands; (iii) COS Shrub classes are particularly present in CLC classes of Shrubs, Pastures as well as in Open spaces with little or no vegetation; (iv) Tree vegetation in COS is also well related to the correspondent CLC classes and allows us to better understand the composition of Forests and Agricultural areas; for example, pure or mix forests of Cork and Holm oak trees are particularly evident in the CLC classes of Agro-forestry areas, Broad-leaved forests, and Non-irrigated arable lands; on the other hand, Eucalyptus (pure or mixed forests) is an important component of the CLC classes of Mixed forests. Complex cultivation patterns. Annual crops associated with permanent crops, and Broad-leaved forests; finally, pure pinus pinaster forests in COS are comprised in the CLC classes of Coniferous forests, Transitional woodland-shrub, and are especially important in Beaches, dunes, sands, where account for 30% of total area; this finding is in good agreement with the presence of pinus forests in the entire central western coast.

It is important to underline that, from 1990 to 2012, the increase in the BA within the RUI is higher (100%) than the increase in the RUI area (70%). This result suggests that other factors, besides the increase of the RUI area, are responsible of the increase of the burnable area within the RUI. In this regard, it is important to take into account some of the specific characteristics of the country, well described in terms of demographic, territory and forest statistics compiled in Feliciano et al. (2015), which can help to understand the most important factors affecting the forest management in Portugal. Forest is nowadays the dominant land use in the country (with more than 35% of total area).
followed by bushes and grasslands (>29%), and agricultural areas (>24%) (Feliciano et al., 2015; FAO, 2018). According to the National Forest Inventory (IFN, 2010), in the 1995-2010 period, four tree species occupied about 85% of total forest area: Eucalyptus (22%-27%), Cork oak (23%-24%), Maritime pine (30%-23%) and, Holm oak (10%-11%). The first three species have the ability to generate land and business income exceeding 50 euros/ha/year (CM, 2015). Most of the forests and wooded lands (~93% in 1995) have non-industrial private owners and there is a high fragmentation of the forest property, particularly evident in the private sector (Mendes et al., 2004). Management practices are also very different and changed significantly in the last years, especially in non-industrial private forest (Novais and Canadas, 2014). According to Feliciano et al. (2015), 1/3 of Eucalyptus area is well managed by the industrial pulp and paper companies, with their own forest management and wildfire prevention/fighting resources, while the remaining area is managed by non-industrial private owners, characterized by different objectives and economic logics. In addition, there is a significant heterogeneity in the spatial distribution of all these characteristics/factors (Baptista and Santos, 2005). Small forest holdings (<10 ha), mainly composed by pine and eucalyptus with low profitability, are much more frequent in the northern and central Portugal, while large properties (>100 ha), essentially of cork oak or a complex and unique agroforestry system of cork oak savanna (“montado”), are predominant in the southern regions of the country. Table 4 of the annex/supplementary material, provides a general description of the main characteristics of forest holdings and forest owners and summarizes the interrelationship between these factors.

Results from our analyses shows that total RUI area increase of about 40% from 1990 to 2012, but in the second half of the investigated period (2000-2006 and 2006-2012) the growth rate of RUI was lower than in the first decade (1990-2000), probably due to a lower growth rate in the process of urbanization of rural areas. It is also important to underline Moreover, the decrease of relative BA within the RUI from 2006 to 2012. These changes could be associated with the consequence of a relative decrease of BA in the recent economic and financial crisis, such as immigration and emigration trends (Cairns, 2011; Ganga et al., 2016; Fonseca and McGarrigle, 2014) and changes in recent plans for territorial spatial planning and protection of forest against forest fires (Mateus and Fernandes, 2014; Parente et al., 2016). At European level, urban sprawl’s complexity and magnitude motivated the European Commission to recommend actions to ameliorate their impacts and to coordinate land use policies, within the European Cohesion Policy 2007—2013 period (CEC, 2006; EEA, 2006). In Portugal these efforts were complemented with national programs and regional plans such as the National Policy and Territorial Management Regional Plan for Territorial Planning (Plano Regional de Ordenamento de Território, PROT) and the Regional Plan for Territorial Planning (Plano de Ordenamento do Território, PROTI), supporting the sustainable development and the environmental landscape quality of NUTS-III areas (Noronha Vaz et al., 2012). However, as far as we know, there is no a specific or general Portuguese legislation about WUI or RUI. It only exists one general mention about RUI in the National Plan to Protect the Forests against Wildfires (CM, 2009). In this plan it is suggested that in order to protect urban-forest interfaces it is necessary to create and maintain external buffer strips (10-100 m) around population clusters, especially in those with the highest fire vulnerability, as well as around parks, industrial polygons, landfills, housing, shipyards, warehouses, and other buildings.

Finally, we firmly believe that the results of this study are sufficiently motivating to promote the development of specific policies and legislation as well as changes in forest and fire management. The increase of the RUI area and particularly of the BA within the RUI clearly suggests the need of improving fire prevention measures and preparedness policies for this interface region. In fact, as indicated by
the Portuguese National Fire Plan 2006 (Oliveira, 2005), the increase of the urban/rural interface, as a consequence of the above-mentioned LULCC, causes this area to be under the use of people not educated for fire and unaware of possible source of ignition. In particular, portuguese forestry/forest managers must prioritize sustainable forest management practices and make brush clearing obligatory. These paradigm shifts make even more sense if one takes into account that the risk of fire is likely to increase in a future climate with a higher frequency of longer and more intense extreme events, such as drought and heat waves.

6 Conclusions

Continental Portugal registered important land cover changes (about 9% of total land area) and an increase of the rural-urban interface increase in the investigated period (1990-2012.). Most significant changes associated to transitions from the following CORINE Land Cover classes: Agricultural areas (35.1%) and Forest and semi-natural areas (15.2%) to Artificial surfaces (including Urban areas); Agricultural areas to Forest and semi-natural areas (7.3%) and vice-versa (6.3%). However, relative net changes in the areas of these main land cover types are appreciable only for Artificial surfaces, representing a substantial increase (of about 50%) in these areas, while Forest and semi-natural areas remains almost constant (0.3%) and Agricultural areas slightly decreased (-4.4%). The spatial distribution of these changes was far from uniform within the territory. Urban sprawl was concentrated in the metropolitan areas of Lisbon and Porto, as well as in central-north and south coastal areas (region of Algarve). Main drivers for the expansion and reconfiguration of urban areas are the prosperity and promoted socioeconomic development, the intense rural abandonment and, in the particular case of Algarve, the development of mass tourist industry. On the other hand, transitions in the south and interior north regions we assisted to a transition to vegetated land use/land cover types occurred mainly in the south and interior north regions.

LULC subclasses with higher, probably caused by deforestation/afforestation and rural abandonment. The CLC classes mainly affected by these changes in the study period were: Forests and Heterogeneous agricultural areas; the increase in Artificial surfaces is due to transitions from these land cover/land use. Vegetated classes with higher burnt area within the RUI in the study period were: Transitional woodland-shrub, the three types of Forests considered in the CLC inventories and, the three sub-levels of Heterogeneous agricultural areas. These findings suggest the needs of extending the notion of wildland-urban interface for Portugal to rural-urban interface, defined as the Forest semi-natural plus Heterogeneous agricultural areas adjacent to Artificial surfaces. Vegetated classes with higher burnt area within RUI in the study period were precisely transitional woodland-shrub, the three types of Forests.

20 Results of the evolution analysis of RUI’s size, burnt areas and burnt areas within the RUI in the fours investigated periods (1990, 2000, 2006, 2012) allow to conclude that, from 1990 to 2012, RUI increased about 70%, burnt area decreased 35% but, nonetheless, burnt area within the RUI increased 100%. These findings underline the need of frequent monitoring and assessment of land use changes and RUI evolution in Portugal, and reinforce the need to focus the attention of forest and fire managers on this highly fire prone region.

As indicated by the Portuguese National Fire Plan 2006 (Oliveira, 2005), the increase of the urban/rural interface, as a consequence of the above-mentioned LULC dynamics, causes this area to be under the use of people “not educated” for fire and unaware of the origin of ignition. There are strong evidences that urban and RUI’s expansion increase fire density and risk (Lampti-Maillet et al., 2011; Fox et al., 2015; Gallardo et al., 2015; Viegas et al., 2015), cost of houses protection from fire (Pellizzaro et al., 2012) and other impacts beyond fire, such as on biodiversity and ecosystems. Areas with higher density of housing, human population and roads tend to exhibit lower population of migrant birds, wolves and other carnivorous as well as species richness of mammals and butterflies, while urban development lead to habitat loss and fragmentation which, in turn, lead to smaller habitat patches and remnant populations (Radeloff et al., 2005).
The conclusions of this study do not prevent but, instead, suggest and encourage more accurate analyses at large scale for characterizing and mapping RUI, using high resolution and precise data (e.g. true houses footprints, road network, census data) to provide practical indications in term of land and fire management. Nevertheless, our study provides precious suggestions as for what is the global distribution and evolution of RUI in Portugal and identifying which regions need to be prioritized in term of RUI monitoring and forest fire protection.

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Annexes/supplementary material

<table>
<thead>
<tr>
<th>Annex/supplementary material</th>
<th>Table 3 – Fraction of COS2007v2.0 classes’ area within CORINE land cover classes.</th>
</tr>
</thead>
<tbody>
<tr>
<td>20</td>
<td>45</td>
</tr>
<tr>
<td>Area</td>
<td>&lt;1 ha</td>
</tr>
<tr>
<td>------</td>
<td>-------</td>
</tr>
<tr>
<td>Forest owners (%)</td>
<td>31%</td>
</tr>
<tr>
<td>Area (%)</td>
<td>10%</td>
</tr>
<tr>
<td>Main tree species</td>
<td>Maritime pine</td>
</tr>
<tr>
<td>Investment</td>
<td>No investment</td>
</tr>
<tr>
<td>Management practices</td>
<td>No active management</td>
</tr>
<tr>
<td>Income</td>
<td>Property-reserve irregular income</td>
</tr>
<tr>
<td>Region</td>
<td>Northern and central</td>
</tr>
</tbody>
</table>

Table 4 - Main characteristics of a sample of forest holdings and forest owners studied by Baptista and Santos (2005). Adapted from Feliciano et al. (2015).