Projecting of wave height and water level on reef-lined coasts due to intensified tropical cyclones and sea level rise in Palau to 2100

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Abstract. Tropical cyclones (TCs) and sea level rise (SLR) cause major problems including beach erosion, saltwater intrusion into groundwater, and damage to infrastructure in coastal areas. The magnitude and extent of damage is predicted to increase as a consequence of future climate change and local factors. Upward reef growth has attracted attention for its role as a natural breakwater able to reduce the risks of natural disasters to coastal communities. However, projections of change in the risk to coastal reefs under conditions of intensified TCs and SLR are poorly quantified. In this study we projected the wave height and water level on Melekeok reef in the Palau Islands by 2100, based on wave simulations under intensified TCs (significant wave height at the outer ocean, SWH = 8.7–11.0 m; significant wave period at the outer ocean, SWP = 13–15 s), SLR (0.24–0.98 m). To understand effects of upward reef growth on a reduction of the wave height and water level, the simulation was conducted for two reef condition scenarios: a degraded reef and a healthy reef. Moreover, analyses of reef growth based on a drillcore enabled an assessment of the coral community and rate of reef production that are necessary to reduce the risk to the coast due to TCs and SLR. According to our calculations under intensified TCs and SLR by 2100, significant wave heights at the reef flat (SWH) will increase from 1.05–1.24 m at present to 2.14 m if reefs are degraded. Similarly, by 2100 the water level at the shoreline (WL) will increase from 0.86–2.10 m at present to 1.19–3.45 m if reefs are degraded. These predicted changes will probably cause beach erosion, saltwater intrusion into groundwater, and damage to infrastructure, because the coral reef is located at 3 m above the present mean sea level. The above findings imply that even if the SWH is decreased by 0.1 m by upward reef growth, it will probably reduce the risks of coastal damages. Our results showed that a healthy reef will reduce a maximum of 0.44 m of the SWH. According to analysis of drillcore, corallomnose Acropora corals will be key to reducing the risks, and 2.6–5.8 g CaCO3/m²/y, equivalent to >8% of coral cover, will be required to protect by 2100. Although the present study focused on Melekeok fringing reef, many coral reefs are in the same situation under conditions of intensified TC and SLR. The results of this study are applicable to other reefs. The use of coral reef growth to reduce disaster risk will be more cost-effective than
Building artificial barriers. Benefits in addition to reducing disaster risk include the ecological services provided by reefs, and marine products and tourism. Research of the type described here will be required to advise policy development directed at disaster prevention for small island nations, and for developing and developed countries.

1 Introduction

Approximately 90 tropical cyclones (TCs, also referred to as hurricanes and typhoons) occur globally every year (Frank and Young, 2007, Senewiratne et al., 2012). TCs cause large waves, storm surges, and torrential rainfall, and can lead to coastal erosion, salinization of coastal soils, and damage to infrastructure (Gourley, 2011a). These negative impacts have major economic costs. For example, the economic cost of TC Pam (category 5 on the Saffir-Simpson scale), which affected Vanuatu in March 2015, exceeded US$449 million, equivalent to 64% of the GDP of Vanuatu (GFDRR, 2016).

Numerical projections indicate that climate change will increase the mean maximum wind speed of TCs (Christiansen et al., 2013). Towards the end of the 21st century, the wind speed and minimum central pressure of the most intense super typhoons in the northwest Pacific Ocean are estimated to attain 85–90 m/s and 860 hPa, respectively (Tsuboki et al., 2015). This will probably increase wave heights and water levels on shores at coastal areas. Additionally, sea level rise (SLR) caused by global warming will probably increase the risk of coastal erosion, flooding, and saltwater intrusion of surface water (Woodroffe et al., 2013). Economic development and population growth are also expected to increase the baseline damages (World Bank and UN, 2010). Therefore, the development of policies for adaptation to TCs and SLR is essential if these projected negative impacts in the near future are to be adequately addressed.

There have been several approaches to reducing the wave heights and water levels, including the construction of sea walls and shelters, developing accurate weather forecasts, and developing TC and SLR warning systems (GFDRR, 2016).

However, small island nations and developing countries will need to develop cost-effective strategies to address the problems associated with these phenomena, including the use of ecosystem services provided by mangroves and coral reefs. For example, mangrove restoration has been demonstrated to attenuate wave height, and reduce wave damage and erosion (Wong et al., 2014). It has also been suggested that coral reefs are highly effective as natural breakwaters. More than 150,000 km of the shoreline in 100 countries and territories is thought to receive some protection from reefs (Burke et al., 2011). Meta-analysis has demonstrated that the entire reef system, from reef slope to reef flat, reduces wave height by an average of 84% (Ferrario et al., 2014). Coral reefs are also habitats for diverse marine organisms and provide various services (e.g., tourism and marine products) that benefit human populations. Additionally, it has been suggested that reefs can be considered to represent self-adapting “green infrastructure” (Benedict and McMahon, 2002) in the ocean that protects against SLR and TCs. The growth of Holocene reefs has been reported to have kept pace with the SLR that occurred in the period 19–6 ka (Montaggioni and Brathwaite, 2009). This suggests that reef growth, as a form of green infrastructure, may be able to respond to SLR and contribute to reducing the wave heights and water levels at shores in the future.
However, approximately 75% of the world's coral reefs are subject to local threats such as coastal development, watershed pollution, and overfishing (Burke et al., 2011). Additionally, global climate change, including global warming and ocean acidification, is expected to have major impacts on coral reefs. Sheppard et al. (2005) indicated that decreasing of bottom roughness due to the coral mortality in Seychelles from 1994 to 2004. Ocean acidification reduces the CaCO₃ saturation rate, reducing the calcification rate of corals, slows coral growth, and can lead to the dissolution of the reef frameworks (Anthony et al., 2008). This process also reduces the roughness of reef topography and consequently the increase of wave energy reaching the shore. This indicates that the degraded reef may cause an increase of the wave heights and water levels at shores in the future. However, few studies consider how the effect of reefs on the wave heights and water levels may change as degraded reefs become healthy. Furthermore, if future reefs are degraded, considering potential supplantation (e.g., by coral transplantation) for healthy reefs will be needed. The target corals and the reef production rate are also poorly understood.

This study focuses on coral reefs in Palau Islands. The Palau Islands are rarely affected by TCs, but two severe TCs (Typhoon Bopha in 2012 and Typhoon Haiyan in 2013) recently impacted the islands. These TCs caused 56%-83% loss of coral cover on the shallow slopes of the eastern reefs (Gouveia et al., 2015), and this had a significant impact on coastal areas (e.g., flooding, erosion, and destruction of buildings). Furthermore, there will probably be an eastward shift in the genesis location of TCs in the northwest Pacific in the near future (2075-2099; Murakami et al., 2011). This implies that Palau Islands that currently experience only infrequent and lower-intensity TCs will probably be affected by more frequent and intense TCs in the near future. Consequently, this study has two main aims: Firstly, we provide a quantitative projection of wave heights and water levels for the healthy reefs and the degraded reefs under intensified TC and SLR conditions by 2100, based on a numerical simulation. Secondly, we estimate the potential of main reef builder and the reef production rate necessary to reduce the coastal risk under the predicted reef degradation, based on analysis of drilled cores.

2 Methods

2.1 Study site

The present study site is located in Malekeok state, at the east central coast of Babalassa, the biggest island in Palau (Figure 1a). Malekeok is an important site because it is the national capital of Palau, housing the national government including the executive, legislative and judiciary branch of the government. Malekeok is an ideal site for this study because it is representative of the east coast of Palau in terms of its closeness to the sea and the threats it faces. Understanding what happen in Malekeok can be applied to other states on the east coast regarding preparedness for the future impacts of climate change. The state consists of reefs, long beaches, mangroves, hills, steep ridges, and rivers. Prior to the contact with foreigners, some of the villages started to increase their influence and power by forming alliances through warfare (Rechbeil and McPhail, 1997), and Malekeok and Koror village became the most powerful villages in the islands. During the
Japanese administration (1919–1945), the settlement of the state had moved to the coastal area from inland. In the present-day, the coastal village is located at -3 m above the present mean sea level (MSL). The Melekeok elementary school and Melekeok state office are also located in the coast. Melekeok reef (7.501°N, 134.660°E) is located on the eastern corner of the atoll. There is no artificial breakwater for ocean waves along the reef. Our survey transect is located near the elementary school (Figure 1b), because the school will probably be a place to evacuate during assumed intensified TCs.

2.2 Impacts of historical TCs and Typhoon Bopha

Since 1951, 19 typhoons passed within 150 km of Melekeok reef in Palau, provided by the Digital Typhoon (http://agora.os-umi.ac.jp/digital-typhoon/index.html.en) based on the Japan Meteorological Agency (JMA) best track data (Table 1). Only 1 severe typhoon passed near this study reef in 1990 (Typhoon Mike), and the impact was limited to the northern reef of Palau (Moragas and Cook, 1995). Prior to Typhoon Bopha that passed south of Palau in December 2 2012, it is suspected that no major typhoons had caused significant damage to coral reefs and coastal areas of Palau for over 60 years.

The minimum pressure of Typhoon Bopha center was 935 hPa and the maximum wind speed was 50 m/s (data obtained by Digital Typhoon: http://agora.os-umi.ac.jp/digital-typhoon/index.html.en). In December 2 2012, the average wind speed was 27 m/s around the study site provided by Windguru (see http://www.windguru.cz). 121 km distance from the track of Typhoon Bopha in November 6-7 in 2013, Typhoon Haiyan passed north of Palau Islands. The minimum pressure of Typhoon Haiyan center was 905-920 hPa (data obtained by Digital Typhoon: http://agora.os-umi.ac.jp/digital-typhoon/index.html.en). In the Northern Hemisphere, TCs rotate in a counterclockwise direction and the most severe conditions are generated in the right semicircle, where the maximum wind speeds and storm surge are localized. Therefore, the average wind speed of Typhoons Bopha was stronger than that of Haiyan around the study site. Similarly, the maximum value of significant wave height offshore in the study area was 8.7 m in December 2 2012 during Typhoon Bopha and 7.7 m in November 7 in 2013 during Typhoon Haiyan (obtained from Windguru, see http://www.windguru.cz).

Therefore, the study area was severely damaged (e.g., the destruction of piers and flooding) by Typhoon Bopha.

2.3 Estimation of wave height and water level

We focused on two parameters: (1) the significant wave height (SWH) at the reef flat, located 400 m from the shore (SWH), and (2) the averaged water level at the shore (WL). SWH was defined as the mean wave height of the highest 33% of waves.

We attempted to find in situ recorded data of ocean wave and water level using underwater loggers and/or radar observational systems at the study site, but there was no in situ observation data for ocean wave and water level at the study site. Therefore, we conducted an estimation of wave height and water level based on the CADMAS-SURF (Super Roller Flume for Computer Aided Design of Marine Structures) wave simulation model (CDIT, 2001). For validation of the model, we compared calculated results with in situ measurements of WL and SWH at the reef flat, which is located 400 m from the
shore, despite under non-TCs conditions. Moreover, we conducted local interviews to obtain reliable information of the impacts of Typhoon Bopha.

The model is a specialized numerical wave tank model used for assessing the threshold of destruction for structures (e.g., sea walls); its use also contributes to coastal management decisions. The governing equation in the model is based on the extended Navier-Stokes equations for a two-dimensional wave field in porous media. The model can reproduce highly non-linear wave profiles against various structures, e.g., impact of a wave breaking sea walls (Isobe et al., 1999). The model can apply to wave deformation (e.g., wave shoaling, wave breaking, wave overtopping, and wave run-up) at coral reefs.

The model has been successfully applied to wave characteristics at coral reefs under TC conditions (Yamashita et al., 2008; Hongo et al., 2012; Nakamura et al., 2014; Watanabe et al., 2016).

To calculate wave characteristics in the model, we input 5 parameters: (1) wave condition, (2) obstacle data (e.g., reef structure and rocks), (3) time control, (4) mesh size, and (5) physical parameters (e.g., density of water) for details of parameters see CDIT (2001). To determine a wave condition, we need to input three parameters: incident significant wave height at the outer ocean (SWH_o); incident significant wave period at the outer ocean (SWP_o); and incident water level at the outer ocean (WL_o). The parameters are discussed below.

SWH_o: Since there is no in situ observation systems for ocean wave and water level offshore or onshore using underwater loggers and/or radar observational systems in Palau Islands. The present day SWH_o value was simulated by using the Global Forecast System (GFS) model at 27 km resolution, provided by Windguru (see http://www.windguru.cz). The values for 4 sites (Maleksoo, Koror, North beach, and West Passage) are provided by the model. The GFS model provided a regular wave, and thus we input a regular wave into the model. In this study, we used the data as a forcing the CADMAS-SURF model. The largest SWH_o value at Maleksoo during Typhoon Bopha was 8.70 m.

Numerical experiments have shown that the maximum wind speeds or TCs in the northwest Pacific will increase by 19% by the late 21st century as a consequence of global warming (Tsuboki et al., 2015). This implies that SWH_o will increase in the future, but will probably vary among study sites as a function of wind speed and the path of TCs. Projecting wind speed depends on future greenhouse gas emissions. Therefore, we assumed that TCs are characterized by a minimum central pressure of 900 hPa. We also assumed that the future maximum SWH_o at Palau reef will be comparable to the TCs that typically affect the Ryukyu Islands (northwest Pacific). These include Typhoons 3nanshan in 2006 and Talim in 2005, which JMA reported max wind speeds and SWH_o of 26-34 m/s and 10.6-11.3 m, respectively (JMA, 2012; see http://www.data.jma.go.jp/gmd/koroyu/adw/wavedata/daily/coastwave.html). Consequently, we assumed that by 2100 the SWH_o will range from 8.70 to 11.0 m.

SWP_o: The SWP_o during Typhoon Bopha was 13.0 s (based on the GFS Windguru model; see http://www.windguru.cz) and was recorded as a peak period (P_o). The empirical P_o/SWP o ratio is approximately 1 (≈ 5
0.95), consequently, we assumed the value of $P_{\text{max}}$ equated to SWP. As an analogy, the SWP, during the severe typhoons Shanshan and Talim in the Ryukyu Islands was 13.0–15.0 s (see JMA: http://www.data.jma.go.jp/gmd/kaiyousub_wave/shiri1/dailycoastwave.html). Therefore, we assumed that the future SWP, at Palau reef will range from 13.0 to 15.0 s. We input the SWP, as a regular wave into the model.

... <snip> WTL: We assumed that the WTL, ranges from 0 to 2.78 m above the present MSL, based on future SLR, tidal ranges, and storm surges. The future SLR is predicted to range from +0.24 m to +0.36 m by 2050, and from +0.44 m to +0.98 m by 2100, based on the Intergovernmental Panel on Climate Change (IPCC) scenarios Representative Concentration Pathway (RCP) 2.6 and RCP 8.5, respectively (Church et al., 2013). The RCP was used for the new climate model simulations carried out under the framework of the Coupled Model Intercomparison Project Phase 5 of the World Climate Research Programme.

For RCP 2.6, the radiative forcing peaks at approximately 3 W/m² before 2100 and then declines (IPCC, 2013). For RCP 8.5, the radiative forcing reaches greater than 8.5 W/m² by 2100 and continues to rise for some amount of time (IPCC, 2013). At the Palau Islands, the tidal range is 1.60 m during spring tides, and the high tide is 0.80 m above MSL. Storm surges lead to extreme SLR when TEs make landfall. We assume that intensified TCs are characterized by a minimum central pressure of 900 hPa, and thus WTL will increase to 1.00 m above MSL as a result of the suction effect of TC.

In the model, we input an obstacle data (i.e., topography). To determine the topography, we established a transect of 2000 m width that extended from 21 m above MSL, at the shore to 269 m water depth in the outer ocean (Figure 1b). The topography along the transect was determined using a topographic map (USGS, 1983) on land, and was measured using an automatic level (NIKON TRIMBLE, AE 7) and an aluminum staff from the shore to the reef crest, and a single beam echo sounder (Honda Electronics, PS 7) on the reef slope at water depths of 0–75 m. As water depths or 75–269 m, the topography was assumed to increase with depth at an angle of 23°. The field survey was conducted in July and September 2015.

In this study, we used two reef condition scenarios. The first was that reefs are healthy and have a growth rate equal to the SLR, and the second was that the reef is degraded and no growth occurs (Figure 1c). The difference between healthy reefs and degraded reefs was expressed as the topographic profile in the model. We assumed that healthy reefs are characterized by a predicted vertical reef growth from the reef crest to the upper reef slope.

In this model, the time step was 0.01 s and the calculation time was 3600 s. We used outputs in the time interval for 1801–3600 s. The horizontal mesh size was set as $\Delta x = 20$ m. The vertical mesh size was $\Delta z = 1–5$ m. We used fine vertical mesh sizes for the land to the reef edge and coarse ones for the outer ocean. The density of water was set at 1025 kg/m³. The porosity of the reef structure is 10%. The parameter of bottom roughness was not supported in the model. The input parameters (e.g., WTL) are given as double figures below decimal point because the future SLR is given as double figures below decimal point (e.g., +0.24 m; Church et al., 2013). Therefore, the calculated values of SWH and WTL are given as rounding at triple figures below decimal point. The data for SWH were calculated using the output from the fore-up crossing method.

You need to add a paragraph describing your experiments listed on Table 3 & 4. For example:

"Sensitivity experiments where conducted to investigate the effects of storm surge at various SLR scenarios and for various TC intensities." You can be even more specific.
2.4 In situ measurement and calculation for model validation

For validation of the CADMAS-SURF model, we measured WL and SWH at the study site despite non-TC conditions. We deployed one water level logger (Onset HOBO) on the reef flat, which is located 400 m from the shore, from October 31 to November 1 2017. The logger sampled at 1 s. The logger was corrected for atmospheric pressure variations using another water level logger (Onset HOBO) deployed on land. We used three sets of 30-min measured data. The sampling details are summarized in Table 2.

We conducted a calculation of WL and SWH based on the CADMAS-SURF model. We assumed that the SWH ranges from 1.20 to 1.30 m, the SWP ranges from 12.0 to 14.0 m, and the WL ranges from 0.50 to 0.90 m above the present MSL, from October 31 to November 1 2017. The data of SWH and SWP were obtained from Windguru (see http://www.windguru.cz). We used a value of SWH on the reef flat, which is located 400 m from the shore. The details of input parameters are summarized in Table 2.

2.5 Estimation of future reef production rate

We estimated the potential future rate of reef production (kg CaCO$_3$/m$^2$/y) using a drill core from the reef crest at Ngerdiu Accent reef in the Palau Islands (Figure 1a). One reef crest core (PL-1; 25 m long) was recovered from Ngerdiu Accent reef (Kayanne et al., 2002). The thickness of the Holocene sequence is 14.5 m long. The Holocene sequence comprised two facies: (1) corymbose Acropora facies, and (2) arborescent Acropora facies (Hongo and Kayanne, 2011). The corymbose Acropora facies is characterized by corymbose and tabular Acropora (e.g., Acropora digitifera). These corals are found on distinct reef crests and upper reef slopes in Palau Islands (Kayanne et al., 2002; Yukitomo et al., 2007). The zone is generally characterized by high-energy waves in water depths less than 7 m (Hongo and Kayanne, 2011). The arborescent Acropora facies is characterized by arborescent Acropora (e.g., Acropora muricata/intermedia complex). These corals occupy the inner reef slope and fore reef slope at water depths of less than 20 m in Palau Islands and other reefs in the present-day Pacific Ocean (Montegiili, 2005; Yukitomo et al., 2007). These corals are interpreted to inhabit low- to moderate-energy wave conditions (Hongo and Kayanne, 2011).

We weighed all samples and measured the density of each facies. Assuming that the reef crest has a homogenous structure, the production rate of the reef crest is given by following Eq. (1):

$$ R = \frac{\rho H}{t} $$

where $R$ (kg CaCO$_3$/m$^2$/y) is the production rate of the reef crest, $\rho$ is the density (kg CaCO$_3$/m$^2$), $H$ (m) is the thickness of the reef crest, and $t$ (y) is the duration of vertical reef formation. We used two reported radiocarbon ages for arborescent Acropora facies (PL-1-79: 8.31 ka, -15.1 m below MSL; PL-1-67: 7.39 ka, -12.0 m below MSL) and four radiocarbon ages for corymbose Acropora facies (PL-1-43: 7.25 ka, -6.8 m below MSL; PL-1-26: 7.15 ka, -4.4 m below MSL; PL-1-8: 6.28 ka, -2.5 m below MSL; PL-1-3: 3.92 ka, -1.8 m below MSL) (Kayanne et al., 2002; Hongo and Kayanne, 2011). We
assumed that the range of upward reef growth rate (i.e., $H_t$) was 3.4–37.1 m/kyr (between samples PL-1-79 and PL-1-67, and samples PL-1-67 and PL-1-43) for orobrissoid Acropora species in response to 10 m/kyr of Holocene SLR. Similarly, we assumed that the upward reef growth rates for the corymbose Acropora species in response to 10 m/kyr, 5 m/kyr, and <5 m/kyr of Holocene SLR were 21.0 m/kyr (between samples PL-1-43 and PL-1-26), 2.2 m/kyr (between samples PL-1-26 and PL-1-8), and 0.3 m/kyr (between samples PL-1-8 and PL-1-3), respectively.

3 Results

3.1 Model validation

Figure 2 shows maximum and minimum values of calculated WL and measured WL on the reef flat at the study site. The mean of calculated WL and measured WL were 0.96 and 0.87 m (Test 1), 0.64 and 0.54 m (Test 2), and 0.56 and 0.54 m (Test 3), respectively. The calculated mean WL on the reef flat slightly exceeded (0.07 m) the measured mean values. Similarly, measured SWH and calculated SWH were 0.11 and 0.18 m (Test 1), 0.26 and 0.11 m (Test 2), and 0.22 and 0.10 m (Test 3), respectively. The difference in mean value of SWH between measurements and calculations was found to be 0.07 m. The results showed good correspondence between measurements and calculation data. Therefore, we adopted the calculation results for this study.

3.2 Wave height and water level at Typhoon Bopha

Maineek reef has distinctly zoned landforms, comprising the reef flat and reef slope (Figure 1c). The reef flat is ~1000 m wide and consists of a shallow lagoon (900 m wide) and a reef crest (100 m wide). The shallow lagoon (~1 m deep) is situated between the shore and the reef crest. During the Typhoon Bopha, local people mentioned that beach erosion and destruction of structure (e.g., the pier and a pavilion, ~3 m above MSL) occurred along the shore at the study site (Figure 3).

Moreover, the road and the ground of elementary school (~2.86 m above MSL) along the shore were flooded and that this was never seen for a past ca. 70 years.

According to our wave simulation, the SWH, was found to rapidly decrease from the upper reef slope to the reef crest (Figure 4). Under present-day TCs (8.70 m SWH, 13.0 m SWP.), the SWH at the reef crest was 2.15 m and the SWH, was 1.05 m (case 1, Table 3). The reef crest dissipated 75% of the SWH. The shallow lagoon dissipated 51% of the remaining wave height at the reef crest. The entire reef dissipated 87.9% of the SWH. Moreover, the SWH, was 1.24 m for storm surge under the present-day TCs (case 15, Table 3) and the entire reef dissipated 85.7% of the SWH.

The WL, was 0.86 m for present-day TCs (case 30, Table 4) and the WL, increased to 2.10 m under storm surge conditions (case 44, Table 4). Moreover, the water level at the shore under present-day TCs (i.e., Typhoon Bopha) reached the elevation of road (~2.86 m above MSL) at the study site (Figure 5).
3.3 Future wave height at the reef flat

The SWH$_i$ was found to increase to a maximum of 2.14 m for degraded reefs and to 1.80 m for healthy reefs under intensified TCs, SLR, and storm surges by 2100 (Table 3). An increase in the intensity of TCs will cause an increase in the SWH$_i$. For example, a SWH$_i$ value of 1.22 m for a healthy reef under present TC conditions (8.70 m SWH$_i$, 13.0 m SWP$_i$) (case 16) will increase to 1.52 m (+24.6%) in 2050 with more intense TCs (10.0 m SWH$_i$, 14.0 m SWP$_i$) (case 18), and increase to 1.66 m (+36.1%) with the most intense TCs (11.0 m SWH$_i$, 15.0 m SWP$_i$) (case 20). Overall, the increase in TC intensity will increase the SWH$_i$ by 38.0 ± 16.0% (mean ± SD, n = 17) for degraded reefs and by 30.7 ± 18.2% (mean ± SD, n = 17) for healthy reefs (Table S1). Moreover, the SLR will cause a slight increase in the SWH$_i$. For example, 1.06 m in SWH$_i$ at a healthy reef under the most intense TCs (11.0 m SWH$_i$, 15.0 m SWP$_i$) and 0.24 m in SLR (case 20) will increase by 0.30 m in SLR to 1.70 m (+2.4%, 0.30 m in SLR; case 21) and to 1.77 m (+6.6%, 0.44 m in SLR; case 27). Consequently, the effect of SLR will increase the SWH$_i$ by 6.5 ± 11.0% (mean ± SD, 21) at degraded reefs and by 5.0 ± 9.2% (mean ± SD, n = 23) at healthy reefs (Table S1).

Furthermore, storm surges (1.00 m) will also increase the SWH$_i$. For example, storm surges will cause an increase in the SWH$_i$ from 1.09 m to 1.35 m (+23.9%) at healthy reefs subject to a TC (8.70 m SWH$_i$, 13.0 m SWP$_i$; between cases 3 and 17). As another example, under the most intense TCs (11.0 m SWH$_i$, 15.0 m SWP$_i$ between cases 6 and 20), storm surges will increase the SWH$_i$ by 1.49 m to 1.66 m (+11.4%). Consequently, storm surges will increase the SWH$_i$ by 20.1 ± 14.9% (mean ± SD, n = 13) at degraded reefs, and by 17.3 ± 14.8% (mean ± SD, n = 14) at healthy reefs (Table S1).

The modeling showed that in all but 6 cases (cases 3, 6, 14, 17, 23, and 26) the SWH$_i$ was reduced by 0.01–0.44 m by upward reef growth by 2100 (Table 3, Figure 6). For example, 0.24 m in upward reef growth caused a 0.24 m reduction in the SWH$_i$ (from 1.45 m for degraded reef to 1.21 m for healthy reef) under more intense TCs (10.0 m SWH$_i$, 14.0 m SWP$_i$) (case 4). This indicates that reef growth enhanced the reduction in wave height from 85.5% at degraded reefs to 87.9% at healthy reef (case 4, Table S1). Similarly, under the most intense TCs (11.0 m SWH$_i$, 15.0 m SWP$_i$), SLR (0.98 m), and storm surge (1.00 m) in 2100 (case 29), 0.98 m in reef growth caused a 0.20 m reduction in the SWH$_i$ (from 2.00 m for the degraded reef to 1.80 m for the healthy reef); thus, the role of the reef as a natural breakwater increased from 81.8% for the degraded reef to 83.6% for the healthy reef (Table S1). Overall, as a result of reef growth, the wave reduction rate increased from 84.7% at degraded reefs to 86.0% at healthy reefs (Table S1).

3.4 Future water level at the shore

The modeling showed that the WL$_i$ will increase from 0.86–2.10 m at present to 1.19–3.45 m at degraded reefs and to 1.24–3.51 m at healthy reefs under intensified TCs, SLR, and storm surges by 2100 (Table 4, Figure 7). An increase in the intensity of TCs will cause an increase in the WL$_i$. For example, a 1.24 m WL$_i$ at a healthy reef under a current modeled TC
(8.70 m SWH, 13.0 m SWP) (case 31) with an increase to 1.55 m (+25.0%) under more intense TCs (10.0 m SWH, 14.0 m SWP) (case 33), and increase to 1.90 m (+53.2%) under the most intense TCs (11.0 m SWH, 15.0 m SWP) (case 35). Overall, the increase in intensity of TCs resulted in an increase in the WL by 22.7 ± 15.3% (mean ± SD, n = 17) for the degraded reef and by 21.4 ± 13.3% (mean ± SD, n = 17) for the healthy reef (Table S2).

The WL will also be increased by SLR. For example, the WL at a degraded reef subjected to a present modeled TC (8.70 m SWH, 13.0 m SWP) increased from 1.19 m with 0.24 m SLR (case 31) to 1.50 m (+26.1%) with 0.44 m SLR (case 37), and to 1.82 m (+52.9%) with 0.74 m SLR (case 38). Overall, SLR increased the WL by 8.7 ± 18.1% (mean ± SD, n = 21) for the degraded reef and by 32.2 ± 37.8% (mean ± SD, n = 23) for the healthy reef (Table S2).

Storm surge also directly increased WL. For example, by 2050, this effect caused an increase in WL from 1.87 m to 2.87 m (+100 m, +53.5%) for the degraded reef and from 1.90 m to 2.86 m (+96 m, +50.5%) for the healthy reef under the most intense TCs (11.0 m SWH, 15.0 m SWP) and SLR (0.24 m) (cases 35 and 49, Table S2). Overall, storm surge significantly increased the SWH by 56.5 ± 17.2% (mean ± SD, n = 13) for the degraded reef and by 59.5 ± 27.7% (mean ± SD, n = 14) for the healthy reef (Table S2).

The difference in WL between degraded and healthy reefs was found to range from only -0.11 to 0.07 m by 2100. For example, by 2050, no difference in WL was found between the degraded and healthy reefs under intense TCs (11.0 m SWH, 15.0 m SWP) and SLR of 0.30 m (case 50). Similarly, a difference of only 0.01 m in WL was found between degraded and healthy reefs under more intense TCs (10.0 m SWH, 14.0 m SWP) (case 55), even with a difference of 0.74 m in upward reef growth.

We found that the reef (+2.86 m above MSL) adjacent to the study site would be flooded in both degraded and healthy reefs scenarios in 7 cases (cases 49, 50, 53, 55, 56, 57, and 58, Table 4) of intensified TCs, SLR, and storm surge. In the worst scenario, by 2100 under the most intense TCs (11.0 m SWH, 15.0 m SWP), 0.98 m SLR, and a storm surge of 1.00 m, the WL was found to increase to 3.45 m for the degraded reef and to 3.51 m for the healthy reef (case 58).

3.5 Potential reef production in mitigating wave risk

The Holocene reef density (p), determined from the PL-1 core, was 720 kg CaCO₃/m³ for arborescent Acropora facies and 590 kg CaCO₃/m³ for corymbose Acropora facies. The estimated reef production rate (R) ranged from 2.4 to 26.7 kg CaCO₃/m²/y for arborescent Acropora facies, and from 0.3 to 14.2 kg CaCO₃/m²/y for corymbose Acropora facies, depending on the upward reef growth rate and the Holocene SLR (Figure 8). The lower part of the reef was composed of both arborescent and corymbose Acropora facies when SLR was 10 m/k.y., whereas the upper part of the reef comprised only corymbose Acropora facies when SRL was <5 m/k.y.

The value of R needed for growth of the Molokail reef to keep pace with future SLR to 2100 was calculated to be 5.3 to 7.1 kg CaCO₃/m²/y for arborescent Acropora facies and 2.6 to 5.8 kg CaCO₃/m²/y for corymbose Acropora facies.
based on the assumption that the future CaCO₃ density and reef growth rate will be equivalent to those of the Holocene reef (Figure 8).

4 Discussion

4.1 Wave height and water level increase in the future

Our results show that present-day Malekula reef is highly effective in dissipating waves, with the reef crest alone reducing the SWHₙ by 75% and the entire reef able to reduce the wave height by 88%. Other reefs (e.g., US Virgin Island, Hawaii, Australia, and Guam) have been reported to reduce wave height by an average of 64% ($n = 10, 51\%-71\%$) at the reef crest, and by an average of 84% ($n = 13, 76\%-89\%$) for the entire reef (Ferrario et al., 2014). Generally, greater wave dissipation efficiency is associated with steep topography from the upper reef slope to the reef crest, because of the rapid decrease in water depth (i.e., shoaling of waves), and wider reef flats are reported to have greater dissipation efficiency (Sheppard et al., 2005). The reef in the present study is characterized by a steep reef topography and a wide reef flat (~1000 m wide).

Our wave calculations show that increasing TC intensity, SLR, and storm surges will cause an increase in SWHₙ by 2100. During future TCs, an increasing wind speed will directly cause increasing SWHₙ. Water tank experiments have shown a positive relationship between SLR and increasing wave height (Takayama et al., 1977). Sheppard et al. (2005) reported a positive relationship between SLR and wave energy density (an increase in SLR of ~0.2 m increased the density by ~100 J/m²) at the Seychelles, with the density being proportional to the wave height squared.

Our result of WL shows that the road along the shore at the study site was flooded during an assumed present TC (i.e., Typhoon Bopha). Our simulation data seems to correspond with the observation by the local people although we could not obtain quantitative data. The present results also show that increasing TC intensity, SLR, and storm surges will cause an increase in WLₙ: of Malekula reef by 2100, with SLR and storm surges directly increasing the WLₙ. Furthermore, an increase in SWHₙ as a result of the increasing intensity of TCs will cause an increase in WLₙ. An increase in WLₙ is likely to be explained by the wave setup. Wave setup occurs if waves break in the reef crest-reef slope zone, the wave thrust decreases as the breaking surge travels shoreward, and consequently the water level rises (Gourlay, 2011a). Laboratory experiments and field observations have generally indicated that wave setup increases with increasing incident wave height and wave period (Nakazato et al., 1994; Gourlay, 2011b). This implies that the occurrence of more intense TCs will cause an increase in wave setup. Furthermore, water levels generally increase with decreasing water depth toward the shore (i.e., wave shoaling).

According to our results, the coastal area at Malekula reef will be at greater risk of damage from large waves and increased water level by 2100. For example, TCs generating large waves typically cause significant beach erosion, as occurred in Tuvalu (Connell, 1999; Sato et al., 2010). Moreover, if storm surges occur, the coastal area at Malekula reef will be flooded (Table 4). This could lead to the destruction of infrastructure, because many buildings (including the elementary
are located at -3 m above the MSL. In addition, saltwater intrusion into groundwater could cause long-term problems for water management, including degrading water quality for drinking and agriculture (Ratzol and Fletcher, 2013).

4.2 Coastal risk reduction through future reef growth

Our results indicate that there is no significant change in the WIL between a degraded reef and a healthy reef. This can be explained by the nature of coral reefs, which are porous structures characterized by a high degree of water permeability. A reef framework has a wide range of porosities from low (where internal cavities have been infilled with marine cements) to high (e.g., a reef framework is mainly composed of branching corals) (Hopkin, 2011). In this study, mean porosity of reef framework is estimated as 10%. This means that sea water permeates through the reef due to porosity, even if the reef is characterized by three-dimensional structures. The above implies that a decreasing of SWH is very important to reduce the risks of coastal damages. Our results indicate that the future WIL will almost reach the elevation of the road at the study site.

The result implies that an increase in wave height of only 0.1 m leads to an increase in risks of substantial coastal damages such as flooding, destruction of constructions (houses and buildings), saltwater intrusion into groundwater, and coastal erosion. Detailed quantity of the damages was beyond the scope of the present study, but flooding mostly occurs within a 1 km wide coastal zone along the shoreline, and 0.33 m of water level rise has little effect on inundation, but 0.66 m of water level rise, results widespread groundwater inundation of the land surface at Oahu Island in Hawaii (Ratzol and Fletcher, 2013), although it is difficult to directly compare a coastal area between the Molokai reef and the results for Oahu Island. In contrast to the results of WIL, the healthier a reef, the greater its effectiveness in reducing SWH, in the future. On average, reef growth resulted in an increase in the reduction rate from SWH, to SWH, of 84.6% to 86.0%, and it reduced SWH, by a maximum of 0.44 m. The reduction is explained by the following three processes. (1) Future coral growth in the reef crest-upper reef slope zone will increase the dissipation of waves breaking as the water depth decreases (Figure 9). The breaking of waves will occur in shallow water when the ratio of wave height to water depth approaches 0.8 (Gourlay et al., 2011.). Based on many field observations at other reefs and the results of wave tank experiments (Takayama et al., 1971; Nakayama et al., 2011), a rapid decrease in water depth at the zone results in an increase in wave breaking. (2) Upward reef growth will increase the reef angle in the wave breaking zone as a result of a rapid decrease in water depth in this zone. This process also results in an increase in wave breaking. (3) With upward reef growth the wave breaking zone will probably migrate from its present location towards the ocean. This process will expand the area of wave height reduction, and consequently wave heights will decrease on the reef flat. Consequently, the above factors emphasize the need for future reef formation and growth to reduce the risk of damage by waves.

However, our results showed that only 6 cases (cases 3, 6, 14, 17, 23, and 26) the SWH was increased to 0.02 m, 0.18 m by upward reef growth by 2100. One possible explanation of an increase in SWH, is a difference in magnitude of infragravity waves between the degraded reef and the healthy reef. Waves propagating onto shallow reefs steepen and break, and while some of the breaking wave energy propagates shoreward as reformed high-frequency waves, the spectral wave
energy shifts into lower frequencies and long-period (infragravity) waves often dominate (Cheriton et al., 2016). Infragravity waves over shallow reef flats have established relationships between the offshore conditions and resulting reef flat characteristics (e.g., complex bathymetry). Increased wave height and water level due to the infragravity waves have been observed for various coral reefs (Nekeza et al., 1994; Cheriton et al., 2016) and have also been demonstrated in laboratory and modelling studies (Nekeza et al., 1994; Reeder and Bricker, 2015; Shimozono et al., 2015). Under normal wave conditions, the effect is not remarkable phenomenon. In contrast, in extreme wave conditions such as tropical cyclones, extreme waves enhance the effect on coral reefs. For these 2 cases, the upward reef growth reduces water depth in the reef crest-upper reef slope zone and it probably enhances a resonant oscillation of water by infragravity waves. Compared to a test case under normal wave conditions (2.0 m SWH, 9.0 s SWP) with case 23 under TC conditions (11.0 m SWH, 15.0 s SWP), the result showed that the SWH was decreased to 0.01 m by upward reef growth (unpublished data). But, the infragravity waves are known to be generated across the coral reef through nonlinear wave interactions and its overall effect remains unclear. Another possible explanation of the increase in SWH is a short of calculation time. We used a calculation data of 1800 s for the analysis of SWH. To better understand the difference in SWH between healthy reefs and degraded reef, a long calculation time will be probably needed. To clearly understand the difference in SWH, between the degraded reef and the healthy reef, it is necessary to obtain the data using other numerical models (e.g., DEIM-3D: De laire, 2017). Moreover, a water tank experiment would be required to investigate the effect of upward reef growth to the difference in SWH.

4.3 Reef production rate for coastal risk reduction

According to the analysis of drill core in this study, a coralline Acropora facies at a high wave energy condition in water depths less than 7 m and an arborescent Acropora facies at a low to moderate wave conditions in water depths less than 20 m contributed to the Holocene reef growth in the Palau Islands (Kayanne et al., 2002; Hange and Kayanne, 2011). The maximum future SLR is predicted to +0.98 m by 2100 (Church et al., 2013). This implies that arborescent Acropora corals will probably be overturned and broken by high wave energy in shallow water depths, and so will not contribute to upward reef formation at the reef crest by 2100. In contrast, coralline corals at the study site will contribute to reef formation by 2100, in response to future SLR. Although the dominant corals at Malekøk reef have yet to be documented, the coralline Acropora facies on reef crests in the Palau Islands is generally composed of A. digitifera, Acropora hyacinthus, and A. millepora (Kayanne et al., 2002; Yamanaka et al., 2007). These coral types are highly resistant to wave action at water depths of 0–7 m, and their preferred habitat (good light penetration and high oxygen concentrations) enables vigorous upward growth.

Our results indicate that if the present MSL increases by 0.44 m (mean value for RCP 2.6) to 0.74 m (mean value for RCP 8.5) by 2100, maintaining reductions in the wave height at the study reef will require 2.6–4.4 kg CaCO3/m²/y to support upward reef growth by the coralline Acropora facies (Figure 9a). Similarly: 5.8 kg CaCO3/m²/y will be required to
maintain wave height reduction by the facies under 0.98 m SLR by 2100 (highest value for RCP 8.5). Field measurements of the reef crest community at the core site following the mass bleaching event in 1998 showed that the calcification rate decreased from 130 to 74 mmol CaCO_3/m^2/day, equivalent to a rate of 4.7–2.7 kg CaCO_3/m^2/yr (Kaneyama et al., 2005). Coinciding with the bleaching event, the coral cover decreased from 8.1% to 1.4% (Kaneyama et al., 2005). Therefore, if the coral cover is -1% in 2100, the corals will keep pace with 0.44 m SLR under RCP 2.6, but >8% coral cover will be needed under RCP 8.5 to reduce wave height and the risk of coastal damage at the study site. However, if mortality of corymbose Acropora occurs at the study reef in the future because of global impacts (particularly elevated sea surface temperature and ocean acidification) and/or local stresses, it will probably cause a decline in reef effectiveness in reducing wave height. Coral calcification is considered to be highly sensitive to elevated sea surface temperature and ocean acidification. Although there is variability in calcification rates among coral species (Pandolfi et al., 2011), corymbose Acropora species (e.g., A. digitifera) are particularly vulnerable to thermal stresses (Loya et al., 2001, Golbuu et al., 2007). The growth of Acropora polyps at Okinawa Island in the Ryukyu Islands was significantly reduced by ocean acidification (Suwa et al., 2010). Local stresses, including sediment discharge, also have a negative impact on the species (Burke et al., 2011, Hongo and Yamano, 2013). In Palau Islands, the loss of mature coral colonies on the eastern reef slopes may have decreased coral recruitment and led to the opening of space for turf algae around the islands after Typhoon Bopha in 2012 and Typhoon Haiyan in 2013 (Gouvea et al., 2015). There was a major decline in juvenile acroporid coral at the reef slope at Melekeok and along the eastern reef slopes in Palau Islands after Typhoon Bopha in 2012 and Typhoon Haiyan in 2013 (Gouvea et al., 2015). Our calculations for case 9 (TC: 8.70 m SWH, 13.0 m SWP; SLR 0.74 m) show that for a healthy reef, 0.74 m/yr of upward growth produced a reduction of 0.23 m in SWH in 2100. If the reef growth rate decreases to 3.7 m/yr, a reduction of 0.05 m in SWH would be expected (unpublished data). Despite the decrease in juvenile acroporid corals at Palau Islands, early successional corals, especially pocilloporid corals, recruited 6 months after Typhoon Haiyan in 2013 (Gouvea et al., 2015). There is no information for upward reef growth by pocilloporid corals in the islands. To understand the role in coastal risk reductions, an estimation of reef production by pocilloporid corals will need to be considered.

4.4 Reduction of global disaster risk based on the health of coral reefs

This study highlights the importance of maintaining reef growth (as a function of coral cover) in the future, to reduce the risk of coastal damage arising from wave action. Therefore, it is necessary to monitor the cover of reef-building corals, the recruitment of coral larvae, and the occurrence of various stressors. For example, the Palau Islands has a Protected Areas Network (PAN), which consists of marine and terrestrial areas established for the protection of important biological habitats. To reduce future risks, warnings derived from monitoring such areas can indicate the need to remove or reduce stressors, and to consider implementing reef restoration efforts (e.g., coral transplantation). Additionally, we recognize that the ground elevation of construction varies with each building. To evacuate the people from the flooding area, an investigation of ground...
elevation at each construction site. The signboard of elevation will be required. Furthermore, to evaluate the impact of hydrodynamic forces at coastal areas in the islands, establishment of in situ observation systems of wave height, wave period, and water level should be considered. For example, establishment of ultrasonic wave-based wave gauges, observation buoys, and radar-based wave meters are recommended to predict accurately the ocean wave heights and periods to alert people of disasters such as flooding during TCs.

More than 150,000 km of shoreline in 100 countries and territories is thought to receive protection from reefs, which reduce wave energy (Burke et al., 2011). More than 100 million people in Southeast Asia live in reef-associated areas (i.e., within 10 km of the coast and within 30 km of a reef), where fringing reefs predominate (Burke et al., 2011). By 2100, this area and its people are likely to be at risk from wave action because of the increasing intensity of TCs, and from SLR and storm surges, and this is likely to have negative economic and social effects. This study focused on Mekelako reef in the Palau Islands, but our results are applicable to other reefs in the Indo-Pacific and Caribbean regions. The natural breakwater formed by reefs is more cost-effective in coastal protection than the construction of artificial defenses (Ferrario et al., 2014).

Inexpensive but effective plans for coastal protection will be needed by small island nations and developing countries. Reef growth is self-adapting to long-term environmental change including SLR, it also provides a habitat for marine organisms and societal benefits including marine products, tourism, and recreation. Further research is needed to develop a policy of disaster risk reduction based on coral reef growth in the Indo-Pacific and Caribbean regions.

5 Conclusion

This study predicted the risk of coastal damage at Mekelako reef in the Palau Islands in the case of intensification of TCs, and increased SLR and storm surges that are likely to occur during the 21st century. Our results, based on wave height and water level using the CADMAS-SURF wave simulation model, and past coral assemblage and reef growth rates estimated from a drillcore, indicate that the present-day reef is highly effective at dissipating incoming waves. However, more intense TCs and increased SLR and storm surges resulting from climate change will increase wave height at the study site reef flat and water level at the shore. This will increase the risk of beach erosion, saltwater intrusion into groundwater, and damage to infrastructure. However, our sedimentological analysis suggests that reef formation by key reef-building corals, including corals like Acropora (e.g., A. digitifera), may respond to future SLR. The upward reef growth will decrease the wave height on the reef flat, and reduce the risk of coastal damage. The use of coral reefs for disaster risk reduction is a cost-effective approach and includes other benefits derived from the various ecological services provided by living reefs. Future research such as that described in this study will be required for designing ecosystem-based disaster risk reduction policies for small island nations and for developing and developed countries alike.

Our research would be useful in predicting wave height and water level on coral reefs in the present climate and in a future climate. However, the research has uncertainties of the results and requires the following improvements. The present study emphasizes that further research is required regarding a short-term variation in sea level during the El Niño of early
1998, mean sea level was ca. 0.20 m lower than normal in Palau Islands (Cadin, 2009). This was quickly followed by the La Niña of late 1998, during this period the mean sea level was 0.35 m above normal (Cadin, 2009). This was a half-meter change in mean sea level over just a few months. Such information will allow us to better understand changes in wave height and water level in the Palau Islands by 2100. Moreover, the CADMAS-SURF wave simulation model can contribute to our projecting of wave height and water level. However, our study showed that there was a few of differences in wave height and water level between in situ measurements and calculations on the reef flat under non-TCs conditions. To accurately validate the model, further in situ measurements of wave height and water level under TCs conditions will be needed. Reef coral is often influenced by internal flows such as diffracted waves due to topographic effects. In order to understand the complex behavior of waves, 3D wave analysis as well as 3D topography measurement will be required. The present study assumed a significant wave height and a significant wave period as a regular wave under TCs, however, coral reefs during TCs are affected by irregular waves. Consequently, it is necessary to set irregular waves for various TCs conditions using the CADMAS-SURF model and/or other 3D wave model. Finally, the input parameters for the CADMAS-SURF model are obtained at 27 km resolution using the GFS model. To precisely estimate SWH and WLw, in situ observed data of wave height, wave period, and water level should be collected.

Supplement

Table S1 Effects of intensification of tropical cyclones, and increased sea level rise and storm surges on wave height on the reef flat at the study site.

Table S2 Effects of intensification of tropical cyclones, and increased sea level rise and storm surges on water level at the shore at the study site.

Author contribution

C. Hongo had the idea and C. Hongo and H. Kurishara designed this research. All authors conducted data analysis, and the writing of the manuscript.

Competing interests

The authors declare that they have no conflict of interest.